

Methane and nitrous oxide emissions from paddy fields in Japan: An assessment of controlling factor using an intensive regional data set

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1 Methane and nitrous oxide emissions from paddy fields in Japan: an
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11
12 Abstract

13 Rice paddy fields, producing a major staple food to support growing world
14 populations, represent a major source of greenhouse gases (GHGs) from agricultural
15 ecosystems. The GHG emissions, mainly as CH₄ and N₂O from paddy ecosystems, are
16 highly sensitive to both environmental and management factors. Yet the identification of
17 specific factors, a fundamental step for GHG inventory and mitigation, is often limited
18 by data availability. Here, we compiled 572 and 174 data on CH₄ and N₂O emissions,

19 respectively, from paddy fields across Japan, which arguably represents the most
20 intensive GHG data set from paddy fields per region. We hypothesized that statistical
21 analyses of the intensive data set allow the identification of key factors and possible
22 mechanisms that have not been fully appreciated in the previous studies.

23 Important environmental factors newly identified for CH₄ emission were soil
24 type and precipitation pattern. The soil emitted CH₄ the most was Histosols (172%
25 higher) and the least was Andosols (32% lower) compared to the other soil types. Our
26 analysis also revealed that the region of severe summer rainfall (southwestern Japan)
27 tended to have higher CH₄ emission. The most critical management-related factor was
28 straw incorporation and its timing had significant impact as previously reported.
29 Specifically, CH₄ emission was 242% and 59% higher by pre-puddling and post-harvest
30 incorporation, respectively. The CH₄ response to straw incorporation had relatively
31 large uncertainty, which partly resulted from the variation in straw mass and soil type
32 (esp. Andosols). In addition, the soils having inherently low CH₄ emission due
33 presumably to more oxidized conditions had significantly higher response to straw
34 incorporation. Organic amendment increased CH₄ by 35%, while water management
35 effect was unclear.

36 We also found that N₂O accounted only for 5.5% of total global warming

37 potential from the paddy fields and was mainly emitted in fallow season (84% of annual
38 emission). The amount of nitrogen fertilizer added, the commonly-used factor to
39 estimate N₂O emission (e.g., IPCC guideline) showed no significant relationship with
40 the N₂O emission in rice growing season, which may be explained by very low level of
41 fertilizer application in Japanese paddy fields (typically < 100 kg ha⁻¹ y⁻¹) compared to
42 other parts of the world.

43 While some of the findings are unique to specific regions (e.g., Andosols), new
44 findings on the factors and potential mechanisms controlling GHG emissions from rice
45 paddy ecosystems would be useful to develop strategies for regional GHG estimate and
46 for modeling biogeochemical cycle in rice paddy ecosystems.

47

48 **Keywords**

49 Greenhouse gas (GHG)

50 Soil type

51 Organic amendment

52 Fallow season

53 Mitigation

54 Regional assessment

55

56 1. Introduction

57 Agricultural ecosystems are the largest contributor (56%) to global anthropogenic
58 non-CO₂ greenhouse gases (GHGs) and paddy rice cultivation represents 9–11% of the
59 agricultural GHG emissions (IPCC, 2014). Paddy ecosystems mainly emit methane
60 (CH₄) and nitrous oxide (N₂O) (Akiyama *et al.*, 2005; Malyan *et al.*, 2016). Reducing
61 the GHGs from paddy ecosystems is therefore an important option to mitigate global
62 warming. To achieve this, identification of the factors controlling the GHG emissions is
63 an essential step.

64 Both environmental conditions and paddy management practices control CH₄ and
65 N₂O emissions. Specifically, these include climate (precipitation and temperature), soil
66 properties (pH, soil organic carbon content, and drainage capacity), organic matter
67 incorporation (straw, manure, and compost), and water management (Yan *et al.*, 2005;
68 Itoh *et al.*, 2011; Malyan *et al.*, 2016). Using data of CH₄ emissions and associated
69 metadata collected from Asian countries, Yan *et al.* (2005) showed that straw and water
70 management were the major factors controlling CH₄ emission. Nitrous oxide emission,
71 on the other hand, depended largely on N input as well as water management during
72 growing season (Zou *et al.*, 2009; Liu *et al.*, 2010). Much less information is available

73 on N₂O emission in fallow season despite that fallow season N₂O emission often
74 exceeds those of growing season (Chen *et al.*, 1997; Zheng *et al.*, 2000; Yao *et al.*,
75 2014).

76 The set of critical factors controlling the GHG emissions may be region-specific
77 because the effects or the strengths of potential factors depend partly on geographic area
78 and political/cultural boundaries that affect rice cultivation practices. In Japan, for
79 example, rice is grown under widely different climate conditions (mean annual
80 temperature: 9 to 23 °C, annual precipitation: 933 to 2548 mm yr⁻¹) (Statistics Japan,
81 2017) and thus climate might be one of the critical factors in Japan. Soil type is another
82 candidate as the paddy soils in Japan develop from contrasting parent materials, namely
83 volcanic ash (Andosols, accounting for 13% of paddy fields), river sediments (mainly
84 Fluvisols), highly weathered, iron-rich loess and paleosols (Acrisols), and peat
85 (Histosols). Interestingly, Andosols appear to emit less CH₄ compared with the other
86 soil types (Yagi and Minami, 1990; Tsuruta, 1997; Minamikawa and Sakai, 2005)
87 despite their higher organic carbon contents. However, no region-wide comparison has
88 been conducted to date. Besides these environmental factors, paddy management
89 practices in Japan may also differ from those in other countries. Typical water
90 management in Japan is midsummer drainage followed by intermittent drainage (GIO *et*

91 *al.*, 2016). However, it often rains in the drainage season, which makes it difficult to
92 control paddy water level and thus the drainage effect on CH₄ emission was marginal in
93 some cases (Itoh *et al.*, 2011). The amount of N input as chemical fertilizer is typically
94 lower in Japan (ca. < 100 kg N ha⁻¹) (Toriyama, 2002; Mishima *et al.*, 2010) than the
95 other Asian countries (e.g., ca. 180–210 kg N ha⁻¹) (Cassman *et al.*, 2002; Chen *et al.*,
96 2014). Previously reported dependency of N₂O on N input and water management may
97 be obscured in paddy systems under much lower N input regime. Single cropping is
98 more commonly practiced in Japan. Fallow season is thus typically much longer (> 0.5
99 year) than multiple cropping systems, which would lead to a greater contribution of
100 fallow period to annual N₂O emission. For the differences in the environment and
101 management as mentioned above, GHG mitigation options developed for one region are
102 not necessarily effective in other regions. Strategy for GHG mitigation should thus be
103 developed after identifying region-specific critical factors.

104 Japan arguably represents a rice-growing region of the world with the most
105 intensive data on GHG emissions. Field experiments have been conducted across many
106 parts of Japan to assess straw incorporation effect on CH₄, which is useful to assess the
107 underlying mechanisms of CH₄ emission. To a limited extent, N₂O emission data is also
108 available across Japanese paddy fields. Yet these data has not been systematically

109 examined. Our working hypothesis was that the intensive data set available for Japanese
110 paddy ecosystems allows us to identify key factors and the underlying possible
111 mechanisms that have not been fully appreciated in the previous studies including the
112 meta-analysis targeted at entire Asia.

113 Two main objectives of current study were (1) to identify critical factors
114 controlling CH₄ and N₂O emissions from Japanese paddy fields using statistical
115 approach, and (2) to assess potential mechanisms accounting for the variability present
116 in CH₄ emission associated with specific management practices (e.g., straw
117 incorporation, water management). First, we conducted a linear regression analysis
118 using the CH₄ emission data and associated metadata from paddy fields reported in the
119 last 27 years in Japan. We paid a special attention to one of the critical factors identified
120 (i.e., soil type) and discussed possible mechanisms behind low CH₄ production in
121 Andosols. Second, we selected the part of the data set obtained from comparative field
122 experiments and examined potential mechanisms controlling the CH₄ emission
123 variability associated with specific management practice (esp. straw incorporation).
124 Third, we assessed if previously reported factors (N input and water management)
125 control N₂O emission during growing season and discussed the importance of fallow
126 season emission.

127

128 2. Materials and Methods

129 2.1. Data set

130 We collected peer-reviewed articles and conference proceedings that reported
131 annual or seasonal CH₄ and/or N₂O emissions in Japanese paddy fields (including
132 field-lysimeter studies) from 1988 to 2014. We searched the articles written in English
133 and Japanese on Web of Science and CINI, respectively, using the following keywords:
134 “Japan”, “paddy”/”rice”, and “greenhouse gas” or ”N₂O (nitrous oxide)” or ”CH₄
135 (methane)”. Double cropping is uncommon in Japan, so these studies were not included
136 in our data set (MAFF
137 (http://www.maff.go.jp/j/tokei/kouhyou/sakumotu/sakkyou_kome/index.html)). In total,
138 we collected 572 CH₄ data from 66 articles for CH₄ (Table A1). Most data was from
139 rice-growing season, but some reported annual CH₄ emission (n = 11). We assumed CH₄
140 emitted in growing season equals to annual CH₄ emission because CH₄ release and
141 uptake in fallow season were marginal (about 1% of those in growing season)
142 (Nishimura *et al.*, 2004; Hasukawa *et al.*, 2013). For N₂O, on the other hand, we
143 collected 138 data from 16 articles for growing seasons, while only 16 data from five
144 articles was found for fallow season and 20 data from seven articles for annual N₂O

145 emission (Table A2).

146

147 2.2. Search for critical factors on CH₄ emission

148 To investigate the factors controlling the variation in CH₄ emission, we estimated
149 the effect of each factor using Eq. 1 (Table 1). The model structure was based on Yan *et*
150 *al.* (2005) with a minor modification. We used categorical variables only, which allowed
151 us to use maximum number of data and to compare the effect of various environmental
152 and management parameters simultaneously in single analysis. The model, thus, did not
153 account for the amount of incorporated organic matters (straw and manure/compost).
154 We also selected the model parameters to fit Japanese rice cropping system. For
155 example, we removed pre-season water status from model parameters because winter
156 flooding was rarely conducted (4 locations). Most sites are under single (only
157 midsummer drainage) or multiple drainages (intermittent drainage with/without
158 midsummer drainage) while the others are under continuous flooding (GIO *et al.*, 2016).
159 We thus removed “deep water” and “rainfed” from water management variables. Straw
160 incorporation is conducted at various timings as fallow period is long in single cropping
161 system. We grouped the timing of straw incorporation into two: post-harvest (till
162 February) and pre-puddling (after March). Spreading straw on soil surface without

163 tillage was done in some cases. We thus included it as one of the parameters in straw
164 management. Application of rice straw and manure compost, anaerobically digested
165 slurry, biogas slurry, liquid cattle waste, and saccharification residues was grouped as
166 one parameter, ‘manure/compost application’.

167 Two environmental factors we considered in the analysis of CH₄ were soil type
168 and climate. We categorized soil types into three: Andosols, Histosols, and Others.
169 “Others” includes lowland soils, gley soils, and yellow soils. These generally
170 correspond to Fluvisols, Gleysols, Acrisols in FAO soil categories, respectively. We
171 differentiated Histosols as it often shows extremely high CH₄ emission. Similarly,
172 Andosols was differentiated as it tends to emit less CH₄. Climate zone was grouped into
173 North, East, and West zones based on the variation in precipitation and temperature
174 regimes following Japan Meteorology Agency (Fig. S1).

175 We fitted the model (Eq. 1) to the observed data set and estimated the effect of
176 each parameter as relative change from a base parameter. In case of water management,
177 for example, “continuous flooding” was selected as a baseline ($M3_{CF} = 0$). Then, the
178 effect of “simple/multiple drainage” ($M3_D$) was expressed as a difference of
179 $\ln(FluxCH_4)$ from the baseline. Using the estimates, we calculated relative emission
180 (referred “response ratios” hereafter). For example, the response ratio of simple/multiple

181 drainage against continuous flooding (RR_w) was calculated as Eq. 2. A parameter with
 182 >1 response ratio indicates enhanced contribution and < 1 response ratio indicates
 183 reduced contribution to CH_4 emission.

184

$$185 \quad \ln(\text{Flux}CH_4)$$

$$186 \quad = E1_i + E2_j + M1_k + M2_l + M3_m \quad (\text{Eq. 1})$$

187

$$188 \quad RR_w$$

$$189 \quad = \exp(E1_i + E2_j + M1_k + M2_l + M3_D) / \exp(E1_i + E2_j + M1_k + M2_l + M3_{CF})$$

$$190 \quad = \exp(M3_D - M3_{CF})$$

$$191 \quad = \exp(M3_D) \quad (\text{Eq. 2})$$

192

193 Table 1. Parameters and variables.

Parameters	Symbol in the Eqs.	Specific categorical variables used
<i>Environment</i>		
Climate zone	$E1_i$	North, East, West
Soil type	$E2_j$	Andosols, Histosols, Others
<i>Management</i>		
Straw incorporation	$M1_k$	Non-incorporation, Post-harvest incorporation, Pre-puddling incorporation, Spreading (no-tillage)
Manure/Compost application	$M2_l$	Application, Non-application
Water management	$M3_m$	Continuous flooding, Simple/multiple drainage

194

195 2.3. Effects of the paddy managements on CH₄

196 We further investigated the effect of specific paddy management (e.g., straw
197 incorporation) by focusing on the data set from comparative field experiments where
198 paired plots had been managed to control confounding environmental and management
199 factors. The numbers of the paired data sets for straw, manure/compost, and water
200 management were 53 (from 13 studies), 33 (from 9 studies), and 43 (from 6 studies),
201 respectively. The relative change in each paired data (CH₄ emission under specific
202 management divided by that without the management) represents the response ratio for
203 each comparative field experiment. We first estimated the average response ratio of each
204 management by the fitted line (intercept = 0) among the whole paired data set and
205 compared with that estimated by Eq. 1. The variability of the response ratios was
206 evaluated by 95% confidence interval of the slope. The significance of the effect was
207 determined by the departure of the slope from one by accounting for its 95% confidence
208 interval. We further investigated the factors causing the response-ratio variability using
209 the same data set. Details are as follows.

210

211 2.3.1. Straw incorporation

212 We investigated the factors causing the response-ratio variability by the following
213 two steps: (1) assessing nonlinearity in the CH₄ flux relationship between straw
214 incorporation ($FluxCH_{4S}$) and no straw incorporation ($FluxCH_{4NS}$) and (2) regressing
215 $FluxCH_{4S}$ against $FluxCH_{4NS}$ with potential factors.

216 For (1), we evaluated the nonlinearity by comparing the
217 residual-mean-standard-errors (RMSE) of fitting curves and lines. Nonlinearity would
218 emerge when the response ratio ($FluxCH_{4S}/FluxCH_{4NS}$) depends on $FluxCH_{4NS}$. For (2),
219 we used straw mass (α) and soil type ($E2_j$) as the potential factors to account for the
220 response-ratio variability (Eq. 3). The response ratio would be higher as the mass of
221 straw (substrates for CH₄ production) increased. The effect of straw mass is clear
222 especially when straw is incorporated just before the start of next cultivation because
223 most straw incorporated still remains. When straw incorporation is conducted in an
224 early fallow season, on the other hand, major portions of straw is decomposed before
225 the next growing season, making the straw mass effect weaker. We explained the
226 interaction between straw mass and the timing of straw incorporation as the term ($T_x \times$
227 $\ln(1+\alpha)$). Soil type could also affect the straw incorporation effects by controlling the
228 soil reduction processes. For example, Andosols containing more electron acceptors

229 (e.g., Fe³⁺) and showing high drainage capacity may have weaker response.

230

231
$$\ln(\text{FluxCH}_4\text{S}) = a \times \ln(\text{FluxCH}_4\text{NS}) + E2_j + T_x \times \ln(1+\alpha) \quad (\text{Eq. 3})$$

232 a : coefficient

233 T_x : post-harvest or pre-puddling (categorical variables)

234 α : straw mass (t ha⁻¹)

235

236 2.3.2. Manure/compost application

237 Types of manure/compost might affect the response-ratio variability depending
238 on their mineralization degree and chemical properties. So, we conducted regression
239 analysis using Eq. 4 which includes manure/compost type (MC_y) as a parameter.
240 Manure, compost, and the other organic materials were applied in spring (before
241 flooding) in all reports.

242

243
$$\ln(\text{FluxCH}_4\text{M}) = \ln(\text{FluxCH}_4\text{NM}) + MC_y \quad (\text{Eq. 4})$$

244 FluxCH_4M : CH₄ flux when manure/compost was applied

245 FluxCH_4NM : CH₄ flux when manure/compost was not applied

246 MC_y : rice straw compost, manure compost, biogas slurry,

247 anaerobically digested slurry, liquid cattle waste, or saccharification
248 residue (categorical variables)

249

250 2.3.3. Water management

251 Number of drainage practice time might affect the response ratios. Frequent
252 drainage would promote soil aeration and reduce CH₄ emission. Thus, we conducted
253 regression analysis using Eq. 5 including drainage time (D_Z) as a parameter.

254

$$255 \ln(\text{FluxCH}_4 D) = \ln(\text{FluxCH}_4 CF) + D_Z \quad (\text{Eq. 5})$$

256 $\text{FluxCH}_4 D$: CH₄ flux under drainage condition

257 $\text{FluxCH}_4 CF$: CH₄ flux under continuous flooding

258 D_Z : Single or multiple (categorical variables)

259

260 2.4. Analyses on N₂O

261 We estimated statistic values of annual and seasonal N₂O emissions using
262 complete data collected. Total global warming potential (GWP) was estimated by CH₄
263 and N₂O emissions because the two GHGs are the major greenhouse gases from paddy
264 fields. The GWPs of CH₄ and N₂O are 34 and 298, respectively, based on IPCC (2013).

265 The N₂O contribution to total GWP was calculated by GWP of N₂O (GWP_{N₂O}) divided
266 by total GWP (GWP_{total}).

267 For growing season, we assessed the effects of the input rate of chemical N
268 fertilizers and soil water management on the N₂O (*FluxN₂O_G*) by fitting Eq. 6 to the
269 collected data. Previous reports showed that there were linear relationships between the
270 N input and N₂O and the relationships (slopes and intercepts) differed depending on
271 water managements (Liu *et al.* 2010). Thus we added the interaction term (*N input* ×
272 *M3_m*) in the model. For fallow season, on the other hand, we could collect only five
273 reports, thus we qualitatively discussed the potential factors.

274

$$275 \quad \ln(\text{FluxN}_2\text{O}_G) = N \text{ input} + M3_m + N \text{ input} \times M3_m \quad (\text{Eq. 6})$$

$$276 \quad N \text{ input: kg N ha}^{-1}$$

277

278 2.5. Regression and statistical analyses

279 All the regression analyses were conducted using linear model (lm) function in R
280 ver. 3.1.2. For estimating upper and lower 95% confidence limits of CH₄ and N₂O
281 emissions, we adopted a bootstrap method (bias-corrected and accelerated method,
282 10000 resamples) by using bootES package of R ver. 3.1.2.

283

284 3. Results

285 3.1. CH₄

286 3.1.1. Effects of environments/managements on CH₄ emission

287 We first assessed the effect of each environmental and management parameter
288 by applying Eq. 1 to the complete data set. Soil type and straw incorporation strongly
289 correlated with CH₄ emission ($P < 0.001$), while the correlations with climate zone,
290 manure/compost application, and water management were insignificant ($P = 0.13$),
291 weak ($P = 0.04$), and insignificant ($P = 0.13$), respectively (Table 2). The statistical
292 model, which uses categorical parameters only, explained 25% of the CH₄ variability.

293

294 Table 2. ANOVA table of Eq. 1 parameters.

Parameters	df	F	P
Intercept	1	1135	0.000
Climate zone	2	2	0.131
Soil type	2	9	0.000
Straw incorporation	3	25	0.000
Manure/compost	1	4	0.036
Water management	1	2	0.130

295

296 Table 3. The effects of environment and management parameters on CH₄.

Parameters	n	Estimate	SD	t	P	Lower	Upper	Response ratio
------------	---	----------	----	---	---	-------	-------	----------------

						95% CL *	95% CL *	mean	Lower 95% CL *	Upper 95% CL *
<i>Intercept</i>	386	4.42	0.14	32.73	0.000	4.16	4.69			
<i>Environments</i>										
<i>Climate zone</i>										
North	165	0.11	0.11	0.93	0.355	-0.12	0.33	1.11	0.89	1.39
East	131	0.00						1.00		
West	90	0.28	0.14	2.01	0.044	0.01	0.56	1.33	1.01	1.74
<i>Soil type</i>										
Others	310	0.00						1.00		
Andosols	68	-0.39	0.14	-2.90	0.004	-0.66	-0.13	0.68	0.52	0.88
Histosols	8	1.00	0.34	2.93	0.004	0.33	1.67	2.72	1.39	5.31
<i>Managements</i>										
<i>Straw incorporation</i>										
None	143	0.00						1.00		
Post-harvest	145	0.46	0.12	3.80	0.000	0.22	0.70	1.59	1.25	2.01
Spreading (No tillage)	25	0.33	0.21	1.54	0.125	-0.09	0.75	1.39	0.91	2.12
Pre-puddling	73	1.23	0.14	8.61	0.000	0.95	1.51	3.42	2.58	4.52
<i>Manure/Compost application</i>										
No	332	0.00						1.00		
Yes	54	0.30	0.14	2.10	0.036	0.02	0.59	1.35	1.02	1.80
<i>Water management</i>										
Continuous flooding	100	0.00						1.00		
Single/Multiple drainage	286	-0.17	0.11	-1.52	0.130	-0.39	0.05	0.84	0.68	1.05

297 * CL: confidence limit

298

299 The effect of soil type was the largest among the environmental parameters.

300 The response ratios of Andosols and Histosols were 0.68 and 2.72, respectively, which

301 indicates that CH₄ emission from Andosols was 32% lower and that from Histosols was

302 172% higher than the other soil types (Table 3). The response ratios (and its inverse

303 value) of the soil types were larger than those of the climate zones (1.11 to 1.33) (Table
304 3). We were not able to find any differences among climate zones except that CH₄ in
305 West was a little larger than East (response ratio = 1.33, $P < 0.05$) (Table 3).

306 Among the paddy managements considered in the model, straw incorporation
307 showed the highest response ratio: 3.42 for pre-puddling incorporation (Table 3). The
308 effect of post-harvest incorporation was a little lower (response ratio = 1.59) (Table 3).
309 The results show that CH₄ was 242% higher when straw was incorporated before
310 puddling in spring than when straw was not incorporated into soil, while 59% higher
311 when straw was incorporated after harvest in autumn or winter. Though the straw
312 incorporation effects were highly significant ($P < 0.0001$), both response ratios had
313 large uncertainty; 95% confidence intervals of the response ratios were 2.58–4.52 and
314 1.25–2.01 for pre-puddling and post-harvest incorporation, respectively (Table 3).
315 Manure/compost application, on the other hand, showed a positive effect on CH₄
316 (response ratio = 1.35), while water managements showed a negative effect (response
317 ratio = 0.84) (Table 3). Thus, manure/compost application increased CH₄ by 35% and
318 drainage decreased CH₄ by 16%. However, these parameters had weak ($P = 0.024$) and
319 insignificant ($P = 0.130$) correlation with CH₄ emission, respectively, due to the large
320 uncertainties relative to the effect (95 % confidence intervals were 1.02–1.80 and 0.68–

321 1.05 for manure/compost and water management, respectively) (Table 3).

322

323 3.1.2. Effects of the paddy managements: analysis using comparative data

324 **3.1.2.1. Straw incorporation**

325 The analysis using paired-field comparison data allowed us to estimate
326 response ratios of straw incorporation. The response ratio was highly variable and
327 ranged from 1 to 10 (among grey solid and dotted lines in Fig. 1). We assessed the
328 average effect of straw incorporation by applying linear regression to the comparative
329 data (blue lines in Fig. 1). The slopes, i.e., average response ratios, were 4.2 and 2.2 for
330 pre-puddling and post-harvest incorporation, respectively. Both ratios were significantly
331 higher than one ($P < 0.001$), confirming the substantial CH₄ increase by straw
332 incorporation found from Eq. 1 using the complete data set. The response ratios were
333 slightly higher than those estimated by Eq. 1 (3.4 and 1.6, respectively) (Table 3).

334 These linear regression lines, however, only weakly accounted for the
335 variability of the data (Fig. 1). As a result, the average response ratios obtained from the
336 regression lines had rather large uncertainties: the 95% confidence intervals were 3.3–
337 5.1 and 1.6–2.7 for pre-puddling and post-harvest incorporation, respectively. Power
338 functions explained the variability better than the linear lines (red lines in Fig. 1). The

339 RMSE of the fitting lines were reduced from 0.55 to 0.38 and from 0.62 to 0.56 for
340 pre-puddling and post-harvest incorporation, respectively (variables were
341 logarithm-converted). Based on the power functions, the response ratio (y/x) was shown
342 as $64x^{-0.62}$ and $7.8x^{-0.28}$ for pre-puddling and post-harvest, respectively (Fig. 1). These
343 results suggest that the response ratio went down as the basal emission (x) increased
344 (Fig. 1). As a result, the variability of the response ratio was large: 3.0–11.9 and 1.2–3.7
345 for pre-puddling and post-harvest incorporation, respectively, when estimated by the
346 nonlinear functions in the range of observed basal CH₄ under non-straw incorporation
347 condition (15–136 and 15–699 kg CH₄ ha⁻¹ for pre-puddling and post-harvest,
348 respectively). These results show that response ratio was not constant but depends on
349 the original paddy conditions that control CH₄ emission when straw was not
350 incorporated.

351 Soil type and mass of incorporated straw partially explained the variability
352 unaccounted for by the nonlinear relationship (red lines in Fig. 1). The regression
353 analysis using Eq. 3 showed that soil types and mass of straw (pre-puddling only) had
354 significant relation with the CH₄ increase ($P < 0.01$) (Table 4). The equation accounted
355 for 76% of the CH₄ variability under straw incorporation condition. The estimate for
356 Andosols was negative (-0.43) and that for Histosols was positive (1.64) (Table 4).

357 These estimates indicate that the straw incorporation effect was smaller in Andosols and
 358 larger in Histosols compared to the other soil types. The effect of straw mass was
 359 different depending on the timing of straw incorporation. The interaction between straw
 360 mass and the timing was significantly positive for pre-puddling incorporation (0.84, $P <$
 361 0.01), while positive yet insignificant for post-harvest incorporation (0.40, $P = 0.16$)
 362 (Table 4). These results suggest that straw incorporation effect was larger as the straw
 363 mass increased and this tendency was stronger for pre-puddling incorporation than for
 364 post-harvest incorporation.

365

366 Table 4. Factors on the response-ratio variability of straw incorporation.

Parameters	n	Estimate	SD	t	P	Lower 95% CL	Upper 95% CL
<i>Intercept</i>	45	2.24	0.61	3.7	0.001	1.02	3.47
<i>ln (FluxCH₄ NS)</i>	45	0.48	0.07	6.4	0.000	0.33	0.63
<i>Soil type</i>							
Others	36	0.00					
Andosols	8	-0.43	0.15	-2.9	0.006	-0.74	-0.13
Histosols	1	1.64	0.37	4.4	0.000	0.89	2.40
<i>Timing × ln (1 + Straw mass)</i>							
Post-harvest	19	0.40	0.28	1.4	0.164	-0.17	0.98
Pre-puddling	26	0.84	0.29	2.9	0.006	0.26	1.43

367

368 3.1.2.2. Manure/compost application

369 We evaluated the effect of manure/compost application on CH₄ using the paired

370 comparison data. The manure/compost application effect was significantly positive. The
371 inclination of the linear regression line, i.e., the average response ratio, was 1.5 (95%
372 confidence interval = 1.2–1.7) (Fig. S2). The response ratio was slightly higher
373 compared to that estimated in section 3.1.1 (1.4) (Table 3). The regression analysis
374 using Eq. 4 showed that types of manure/compost was insignificantly related with the
375 CH₄ increase by manure/compost application and thus little related with the
376 response-ratio variability ($P = 0.86$).

377

378 **3.1.2.3. Water management**

379 The effect of water management was evaluated using the paired comparison
380 data. The average response ratio was 0.61 and significantly lower than one (95%
381 confidence interval = 0.53–0.68) (Fig. 2), showing that the water drainage significantly
382 reduced CH₄ emission. The response ratio was lower than that estimated in section 3.1.1
383 (response ratio = 0.84, insignificant) (Table 3). The CH₄ reduction by water drainage
384 was not related with the number of drainage practice times (single or multiple) (using
385 Eq. 5, $P = 0.27$).

386

387 3.1.3. Methane emission as categorized by soil type and paddy management

388 The consistent finding among the series of regression and statistical analyses
 389 was that soil type and the paddy management (straw incorporation, manure/compost
 390 application, and water management) had significant effects on CH₄ emission. We thus
 391 summarized the collected data as follows (Table 5).

392

393 Table 5. Methane emission as categorized by soil type and paddy management.

Paddy managements	Article n	Data n	Statistics								
			Mean	Median	SD	Skewness	Min	Max	Lower 95% CL*	Upper 95% CL*	
<i>Multiple/Simple drainage</i>											
<i>Andosols</i>											
None	10	28	67	63	46	1.5	-1	219	53	87	
Post-harvest straw incorporation	6	20	111	72	142	2.8	-1	615	69	210	
Pre-puddling straw incorporation	1	1	71	71	na	na	71	71	na	na	
Manure/compost application	3	11	94	78	56	0.8	19	206	67	132	
<i>Histosols</i>											
None	1	3	101	78	78	1.2	38	188	38	151	
Post-harvest straw incorporation	2	7	227	226	102	1.0	108	415	169	315	
Pre-puddling straw incorporation	0	na	na	na	na	na	na	na	na	na	
Manure/compost application	0	0	na	na	na	na	na	na	na	na	
<i>Others</i>											
None	16	59	111	80	95	1.9	3	424	90	139	
Post-harvest straw incorporation	18	119	187	156	138	1.3	10	684	164	214	
Pre-puddling straw incorporation	14	50	400	322	236	1.6	116	1189	344	476	
Manure/compost application	8	23	170	116	102	0.5	29	394	131	213	
<i>Continuous flooding</i>											
<i>Andosols</i>											
None	3	8	84	67	79	1.6	12	252	47	159	
Post-harvest straw incorporation	3	6	115	100	63	1.3	56	227	79	178	
Pre-puddling straw incorporation	1	1	100	100	na	na	100	100	na	na	
Manure/compost application	0	0	na	na	na	na	na	na	na	na	
<i>Histosols</i>											

	None	2	3	498	283	409	1.7	241	969	241	740
Post-harvest straw incorporation		2	2	898	898	701	na	402	1393	402	898
Pre-puddling straw incorporation		0	0	na	na	na	na	na	na	na	na
Manure/compost application		0	0	na	na	na	na	na	na	na	na
							<i>Others</i>				
	None	7	24	122	84	125	1.3	2	411	81	181
Post-harvest straw incorporation		4	16	191	127	146	0.6	34	449	130	270
Pre-puddling straw incorporation		9	24	354	351	218	1.3	80	1044	284	462
Manure/compost application		1	8	375	304	365	1.1	2	1090	186	680

394

395 3.2. N₂O

396 We assessed N₂O emission in growing season and fallow season, separately. The
397 N₂O emission in fallow season was larger than growing season. Mean values were 0.3
398 kg N₂O ha⁻¹ and 1.6 kg N₂O ha⁻¹ for growing and fallow season, respectively (Table 6).
399 Some N₂O emissions were extremely high in fallow season (circled red symbols in Fig.
400 3) (Ishibashi *et al.*, 2007; Ishibashi *et al.*, 2009). Even after removing the high data
401 points that gave the mean value of 0.5 kg N₂O-N ha⁻¹, the emission was still
402 significantly higher than that of growing season ($P < 0.05$).

403 We conducted a liner regression analysis using Eq. 6 to investigate the factors
404 on N₂O in growing season. We found that the amount of chemical N fertilizer added,
405 water management and those interaction were all insignificantly correlated with the N₂O
406 emission ($P = 0.24, 0.39, \text{ and } 0.53$, respectively) (Table 7, Fig. 3). In fallow season, we
407 were able to collect only a few data (Table 6, red symbols in Fig. 3). Higher N₂O was

408 observed in the following cases: the use of controlled-release fertilizers (LP100 and
 409 140), spreading straw on soil surface after harvest, and no tillage (circled red symbols in
 410 Fig 3).

411

412 Table 6. N₂O emission and its contribution to total GWP.

		Article n	Data n	Mean	Median	SD	Skewness	Min	Max	Lower 95% CL	Upper 95% CL
N ₂ O (kg N ₂ O-N ha ⁻¹)	Growing season	16	138	0.3	0.3	0.4	3.3	-0.5	2.7	0.2	0.3
	Fallow season	5	16	1.6	0.8	1.8	1.4	0.1	5.9	1.0	2.8
	Annual	7	20	1.6	0.9	1.7	1.6	0.1	6.1	1.0	2.6
GWP _{N₂O} /GWP _{total} (%)	Growing season	16	133	4.6	0.7	15.1	5.8	-7.7	115.1	2.7	8.3
	Annual	6	20	5.5	3.9	5.1	1.4	0.2	20.7	3.7	8.2

413

414 Table 7. Relationship between the parameters and N₂O in growing season.

Parameters	n	Estimate	SD	t	P
<i>Intercept</i>	122	-1.3	0.6	-2.1	0.04
<i>N input</i>		0.0093	0.0079	1.2	0.24
<i>Water management</i>					
Continuous flooding	19	0.00			
Simple/Multiple drainage	103	0.58	0.67	0.86	0.39
<i>Water management × N input</i>					
Continuous flooding × N input	19	0.0000			
Simple/Multiple drainage × N input	103	-0.0053	0.0085	-0.63	0.53

415

416 3.3. Contribution of CH₄ and N₂O to GWP

417 Global warming potential mainly depended on CH₄ with minor contribution from
 418 N₂O. On annual basis, the ratio of GWP_{N₂O} to GWP_{total} was 5.5% on average and 3.9%

419 as a median (Table 6). The N₂O contribution was based on six references including
420 Ishibashi *et al.*, 2007 and 2009 which showed extremely large N₂O emissions (Fig. 3,
421 S3). When we removed this data, GWP_{N_2O}/GWP_{total} was reduced to 4.8% (mean) and
422 2.0% (median). The N₂O contribution was much less in growing season: 4.6% as a
423 mean and 0.7% as a median (Table 6). The large difference between mean and median
424 resulted from the overestimation of GWP_{N_2O}/GWP_{total} when CH₄ emission (and thus
425 total GWP) was extremely low, for example, in Andosols without straw incorporation
426 (Fig. S3).

427

428 4. Discussion

429 4.1. Environmental factors controlling CH₄ emission

430 The analyses using the CH₄ emission data and associated metadata from Japanese
431 paddy fields showed that soil types and straw incorporation were the most influential
432 factors accounting for its variation. In addition to the fact that these two factors strongly
433 correlated with CH₄ (Table 2), the response ratio analysis showed that the specific
434 parameters strongly controlling CH₄ variability were Histosols among the environment
435 parameters and pre-puddling straw incorporation among the management parameters
436 (Table 3). The significant effect of straw incorporation on CH₄ emission is well known

437 (Feng *et al.*, 2013; Liu *et al.*, 2014). However, no previous studies using large data sets
438 have shown that soil type can be one of the most critical factors explaining the CH₄
439 emission.

440 Soil type had a substantial effect on CH₄ emission. While Andosols and
441 Histosols are rather minor soil types in Japan (13% and 6% of total rice cultivation area,
442 respectively), their effects per area were large (Table 3). While 2.7-fold increase was
443 observed, more data is needed to estimate more accurate response ratio for Histosols as
444 our data set contains only eight data points (Table 3). Much higher number of data
445 points for Andosols across Japan (68 points, Table 3) allowed us to conclude that CH₄
446 emission from paddy fields is significantly lower in Andosols compared to the other soil
447 types for the first time. By multiplying the distribution area by the response ratio for
448 each soil type, we estimated the relative contribution of soil type. The CH₄ emission
449 from paddy was reduced by 9% due to the presence of Andosols and increased by 16%
450 for Histosols, highlighting the importance of soil type to improve the process-based
451 model used in Tier III methodology in Japan (Hayano *et al.*, 2013; Katayanagi *et al.*,
452 2016).

453 How do Andosols achieve reduced CH₄ emission despite significantly higher
454 soil organic C contents compared to the other soil types? The mean of total C (TC)

455 content under Andosols was 50 mg kg^{-1} , which is more than 2-fold higher than that
456 under the other soil types (22 mg kg^{-1}). Soil pH and TC content have been identified as
457 the soil factors controlling CH_4 emission from paddy fields (Yan *et al.*, 2005). When we
458 conducted a regression analysis by adding pH and TC to Eq. 1, soil pH insignificantly
459 related with CH_4 ($P = 0.56$) and soil TC content significantly and positively related with
460 CH_4 ($P < 0.001$) (Table S2). The positive TC effect suggests that CH_4 emission from
461 Andosols would be higher in contrast with our observation. Thus, these factors (pH and
462 TC) had little effect and other soil properties are likely to be responsible for the reduced
463 CH_4 emission from Andosols. The response of CH_4 to straw incorporation (i.e.,
464 sensitivity to addition of the substrates) was low in Andosols (Table 4), thus substrate
465 limitation may not be a main factor. A possible mechanism is Andosol's capacity to
466 promote or maintain high redox potential. Rapid water drainage and unique mineralogy
467 have been pointed out as possible factors contributing to the reduced CH_4 of Andosols
468 (Sass and Fisher, 1997; Cheng *et al.*, 2007). Andosols are characterized by highly
469 porous structure and thus show high hydraulic conductivity under paddy management
470 (Nanzyo *et al.*, 1993), which leads to the maintenance of soil redox condition high
471 enough to inhibit methanogenesis (Yagi and Minami, 1990; Minamikawa and Sakai,
472 2005). Andosols often contain large amounts of oxalate-extractable iron oxides mostly

473 composed of ferrihydrite (Makino, 2001; Imaya *et al.*, 2007; Suda *et al.*, 2012). The
474 oxidized forms of iron directly and indirectly inhibit methanogenesis (Kato *et al.*, 2012;
475 Zhou *et al.*, 2014). However, specific mechanisms behind the reduced CH₄ emission
476 from Andosols remain unclear and thus deserve further investigation as the mechanistic
477 understanding may be useful when developing mitigation strategies and improving
478 GHG estimation models. For instance, we may gain some insights on how biochar
479 application affects CH₄ emission because high porosity and TC contents of Andosols are
480 partly attributable to the abundance of charred materials (Shindo *et al.*, 2004).

481 Climate control on the CH₄ emission from paddy fields became apparent after
482 the analysis with the zonation which was based largely on precipitation seasonality. No
483 significant difference among climate zones was found except that CH₄ emission from
484 West zone was slightly higher than that from East zone (Table 3). Further regression
485 analysis by breaking down West zone into northern and southern sections revealed
486 significantly high CH₄ emission from the southern section (i.e., south western Japan)
487 with the response ratio of 1.68 ($P < 0.0001$, data not shown). This region experiences
488 high rainfall in summer (Fig. S1) and thus the soils tend to be under reduced condition
489 even after paddy drainage, which may account for the higher CH₄ emission (Ishibashi *et*
490 *al.*, 1997; Ishibashi *et al.*, 2001; Sasaki *et al.*, 2002). The heavy summer rainfall

491 together with high temperature might synergistically promote methanogenesis in south
492 western area, leading to the higher estimate of CH₄ emission from West zone. In North
493 zone, on the other hand, temperature and precipitation might affect in opposite direction.
494 Lower temperature in this zone reduces methanogen activity, while snow melt leading
495 to reduced redox condition at early stage of growing season would promote
496 methanogenesis (Shiratori *et al.*, 2007). These opposing effects of temperature and
497 precipitation are likely to result in the insignificant difference between North and East
498 zones. These results suggest that zonation based on seasonal precipitation patterns may
499 be an effective approach when assessing CH₄ emission from rice paddy.

500 We identified several critical controls on CH₄ emission using the large data set.
501 However, the parameters used for this analysis accounted for only a quarter of the CH₄
502 variability. This is partly because we did not account for interactive effects. We thus
503 conducted further analysis on management effect by focusing on comparative field
504 experiments.

505

506 4.2. Management factors controlling CH₄ emission

507 How straw incorporation affected CH₄ emission deserves detail examination as
508 our analyses revealed that this was the management practice with the highest sensitivity

509 in Japanese paddy fields. The CH₄ response ratios were high: 3.4 and 1.6 for
510 pre-puddling and post-harvest incorporation, respectively (Table 3). These response
511 ratios were comparable to previous reports: 3.1 and 1.8, respectively, at 6t ha⁻¹ straw
512 incorporation (Yan *et al.*, 2005) and 1.4–1.8 for post-harvest incorporation (GIO *et al.*,
513 2014) (Table 3). The analysis using the paired data set showed higher response ratio
514 compared with those estimated by Eq. 1 (Table 3): 4.2 for pre-puddling and 2.2 for
515 post-harvest straw incorporation. The differences in the estimated response ratios
516 between the two data sets used might be attributable to the bias present in the paired
517 comparison data. The basal CH₄ (CH₄ without straw incorporation) was in a lower range
518 where the response ratio was higher (Fig. 1, Table 5), which resulted in the higher
519 average response ratio. Nevertheless, the results from both analyses consistently showed
520 high response ratios for straw incorporation, confirming its strong effect on CH₄
521 emission.

522 No previous studies showed higher response ratio (y/x in Fig. 1) under lower
523 basal CH₄ emission (x in Fig.1) to our limited knowledge. This novel pattern is likely to
524 be controlled by the change in soil redox status by straw incorporation. Methane
525 production is enhanced under strongly reduced condition (Malyan *et al.*, 2016). Thus,
526 the high response ratio found here implies that straw incorporation led to strong

527 reduction. The change in soil redox status can be explained by a simple model on the
528 relative volumetric change in soil responsible for methanogenesis (referred
529 'methanogenic zone' hereafter) (Fig. 4). This conceptual model assumes that (1) the
530 volume of methanogenic zone without straw incorporation is formed by the oxidation of
531 soil organic matter (V_s), (2) the decomposition of incorporated straw creates
532 methanogenic zone in soil (V_{rs}), (3) other factors that may affect methanogenic zone
533 volume (e.g., straw mass, climate condition, water and manure management, and etc.)
534 are constant, and (4) both soil organic matter and incorporated straw that leads to the
535 formation of methanogenic zone are randomly distributed. With these assumptions, total
536 methanogenic zone volume after straw incorporation (V_{s+rs}) can be expressed as
537 $(1-V_{rs})V_s + V_{rs}$ (solid line in Fig. 4) where V_{rs} was a constant. The model shows that
538 soils with lower V_s (under less reduced condition) would have more room for the
539 formation of methanogenic zone by rice straw incorporation (black arrow in Fig. 4),
540 while vice versa in soils with higher V_s (grey arrow in Fig. 4). The pattern of V_{s+rs} was
541 quite similar with Fig. 1. Thus, assuming that the basal CH_4 emission occurred in
542 parallel with V_s , the pattern in the response ratio (Fig. 1) can be explained by the
543 changes in the redox status (Fig. 4). The V_{s+rs} controls CH_4 production associated with
544 rice roots (e.g., root exudates and dead roots), one of major sources of CH_4 in paddy

545 fields (Watanabe *et al.*, 1999; Tokida *et al.*, 2011; Yuan *et al.*, 2012), assuming that root
546 randomly grows within straw mixed soil where methanogenic zone was V_{s+rs} . The CH_4
547 production using straw, the other major source of CH_4 , on the other hand, might depend
548 on V_s and V_{rs} rather than V_{s+rs} . Straw-derived CH_4 is increased as V_s was larger when
549 straw was mixed with soil where methanogenic zone was V_s , and after V_{s+rs} was
550 established after the straw incorporation, straw-derived CH_4 depends on V_{rs} . Then the
551 pattern of straw-derived CH_4 can be more parallel with 1:1 line in Fig. 1 and 4. Based
552 on this idea, large response at lower basal emission in Fig. 1 might mainly be attributed
553 to the response of root-derived CH_4 . At present, there are no reports providing the
554 relationship between original paddy condition and substrate-based CH_4 response. These
555 processes can be quantitatively examined by conducting comparative experiments with
556 isotope-labeling approach using different soil types, which would improve simulation
557 models of C cycling in paddy fields.

558 The mass of straw incorporated and soil type can also affect the CH_4 response
559 in addition to the methanogenic volume control. Larger straw mass led to higher CH_4
560 response (Table 4) because straw decomposition consumes oxygen and straw itself is a
561 substrate for methanogenesis. Lower response of Andosols compared to the other soil
562 types (Table 4) may be attributable to soil physical and mineralogical characteristics as

563 discussed above. These findings imply that, for more accurate CH₄ estimation, grouping
564 based on soil type, straw mass, and the methanogenic condition of original soil (prior to
565 straw incorporation) might be effective.

566 Manure/compost application also affected CH₄, though the effect was not as
567 large as straw incorporation. These application practice increased CH₄ emission by 35%
568 using complete data for Eq. 1 and by 46% based on the paired comparison data (Table 3,
569 Fig. S2). The slightly higher response ratio of the latter may be attributable to the timing
570 of manure/compost application (i.e., spring just before the next cultivation). In some
571 data used for the regression analysis (Eq. 1), manure/compost was applied in fall and
572 the decomposition of manure/compost for several months before next cultivation might
573 have reduced the CH₄ response. The CH₄ response was not significantly different
574 among manure/compost types based on the regression analysis using Eq. 4 (Fig. S2,
575 section 3.1.2.2). Quality parameter (e.g., C:N) might be more useful in categorizing the
576 effects of manure/compost application because the decomposition rate of organic matter
577 is strongly controlled by C:N ratios (Katayanagi *et al.*, 2016).

578 Water drainage management can significantly reduce CH₄ emission as long as
579 we can sufficiently control soil water condition especially during midsummer drainage
580 season. The comparative data estimated significant and large water drainage effect – the

581 response ratio of 0.61 (Fig. 2). The response ratio was comparable to the previous report
582 on Asia (0.60 and 0.52 for single and multiple drainages, respectively) (Yan *et al.*, 2005).
583 However, we were unable to detect such large effect using the complete data (Table 3).
584 In the comparative experiments, researchers usually strictly controlled drainages to
585 detect differences among the water managements. On the other hand, the other articles
586 not focusing on water managements, soils could be insufficiently aerated (incomplete
587 drainage). We found little difference in single (only midsummer drainage) and multiple
588 (intermittent drainage with/without midsummer drainage) drainage effects (Fig. 2),
589 suggesting that sufficient soil aeration during midsummer drainage is critical for CH₄
590 reduction. It often rains in midsummer in Japan (esp. July, Fig. S1) and the precipitation
591 during this season strongly limit the drainage effects (Ishibashi *et al.*, 1997; Ishibashi *et*
592 *al.*, 2001; Sasaki *et al.*, 2002; Itoh *et al.*, 2011). Thus, the CH₄ reduction by 39% (Fig. 2)
593 may be expected only under no heavy rain and/or sufficient drainage in midsummer.

594

595 4.3. Factors controlling N₂O emission and its relative contribution to GWP

596 Fallow season was the critical period for N₂O emission from Japanese rice fields.
597 N₂O emission during fallow periods (7-9 months) was roughly 5-fold higher than that
598 from growing seasons (Table 6). While the contribution of annual N₂O to total GWP

599 was small compared with CH₄ (5.5% by mean and 3.9% by median) (Table 6), the N₂O
600 contribution showed a large uncertainty (95% confidence interval = 3.7–8.2%). The
601 high uncertainty stems partly from the limited availability of data set (six references)
602 that includes a study with extremely high N₂O in fallow season (Ishibashi *et al.*, 2007;
603 Ishibashi *et al.*, 2009). The paddy field examined by Ishibashi *et al.*, 2007 and 2009 has
604 following characteristics: the use of controlled-release fertilizer, the return of straw on
605 soil surface after harvest, and the lack of tillage in fallow season. Any of these practices
606 could increase N₂O emission during fallow period because N derived from the
607 controlled-release fertilizer and the straw left on soil surface can be the source for N₂O.
608 No-tillage tends to conserve soil water, which could promote denitrification and thus
609 N₂O emission (Linn and Doran, 1984). However, we need more information on fallow
610 N₂O emission to reveal the most critical factors.

611 For growing season, on the other hand, several factors (N input and water
612 management) have been identified from the paddy fields in other countries. These
613 factors are, however, only marginally contributing to the N₂O in Japan (Table 7). The
614 results are likely to be attributable to the difference in N fertilizer input (mostly up to
615 100 kg N ha⁻¹ in Japan, Fig. 3). The lower N input led to smaller N₂O emission and its
616 variability so that we were unable to detect the effect of these parameters (Fig. 3, Table

617 7). Reduced N input and enhanced N use efficiency are well in line with the future
618 direction to reduce water pollution as well as N₂O emission. Thus, the emission factor
619 approach (N₂O/N input) used following IPCC guideline (2014) may not be necessary in
620 case of Japanese paddy fields.

621

622 5. Concluding remarks

623 We compiled the studies and reports of CH₄ emission from paddy fields in Japan.
624 The data set consists of 572 data points from paddy systems under a wide range of
625 climatic, edaphic, and management conditions. We also collected 33–53 paired data
626 points from field comparison studies that were designed to assess specific management
627 effects (straw, manure/compost, and water management). Given the long history of rice
628 paddy research in Japan, the data set compiled arguably represents the most intensive
629 paddy CH₄ data set per region in the world.

630 The complete data set was analyzed by the statistical model only with selected
631 categorical variables to maximize the use of data points, which allowed us to identify
632 the two most critical factors (soil type and straw incorporation) controlling CH₄
633 emission from Japanese paddy ecosystems. While this model accounted only for a
634 quarter of the CH₄ variability, the second analysis using the paired experiment data

635 confirmed this finding and provided further insights on the controls of CH₄ emission.
636 Specifically, we found the importance of the timing and the amount of straw
637 incorporation as well as original soil conditions including soil type. These factors may
638 account for the residual CH₄ variability unexplained by the first analysis. While less
639 data points were available for N₂O, we found N₂O emission was significantly greater in
640 fallow season than in growing season.

641 While some of the controlling factors identified are more related to the unique
642 condition of the selected region (e.g., soil type: Andosols), other findings are likely to
643 be applicable to wider regions. For instance, we found that CH₄ increase upon straw
644 incorporation was greater for the soils having low basal CH₄ emission due presumably
645 to greater increase in the soil volume that was strongly reduced. We thus proposed a
646 conceptual model to account for the non-linear behavior of CH₄ response to straw
647 addition. While further validation of the model as well as experimental work are
648 required to fully understand how straw incorporation contributes to CH₄ emission, our
649 findings would be useful when assessing the management effect on CH₄ in other regions.
650 In summary, the findings from the initial categorical data analysis using the large data
651 set and those from the second regression analysis using smaller yet well-controlled
652 experimental data set were complementary with each other. The new findings on the

653 factors and potential mechanisms controlling GHG emissions from rice paddy
654 ecosystems may be useful to develop strategies for regional GHG estimate and for
655 modeling biogeochemical cycle in rice paddy ecosystems.

656

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660

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816

817 **Figure caption**

818

819 Fig. 1. Effect of straw incorporation on CH₄ emission

820 The blue lines are linear approximation lines for post-harvest (solid, $y = 2.2x$) and
821 pre-puddling straw incorporation (dotted, $y = 4.2x$). The red lines are exponent
822 approximation lines for post-harvest (solid, $y = 7.8x^{0.72}$) and pre-puddling straw
823 incorporation (dotted, $y = 64x^{0.38}$). The gray solid and dotted lines are 1:1 and 1:10 lines,
824 respectively.

825

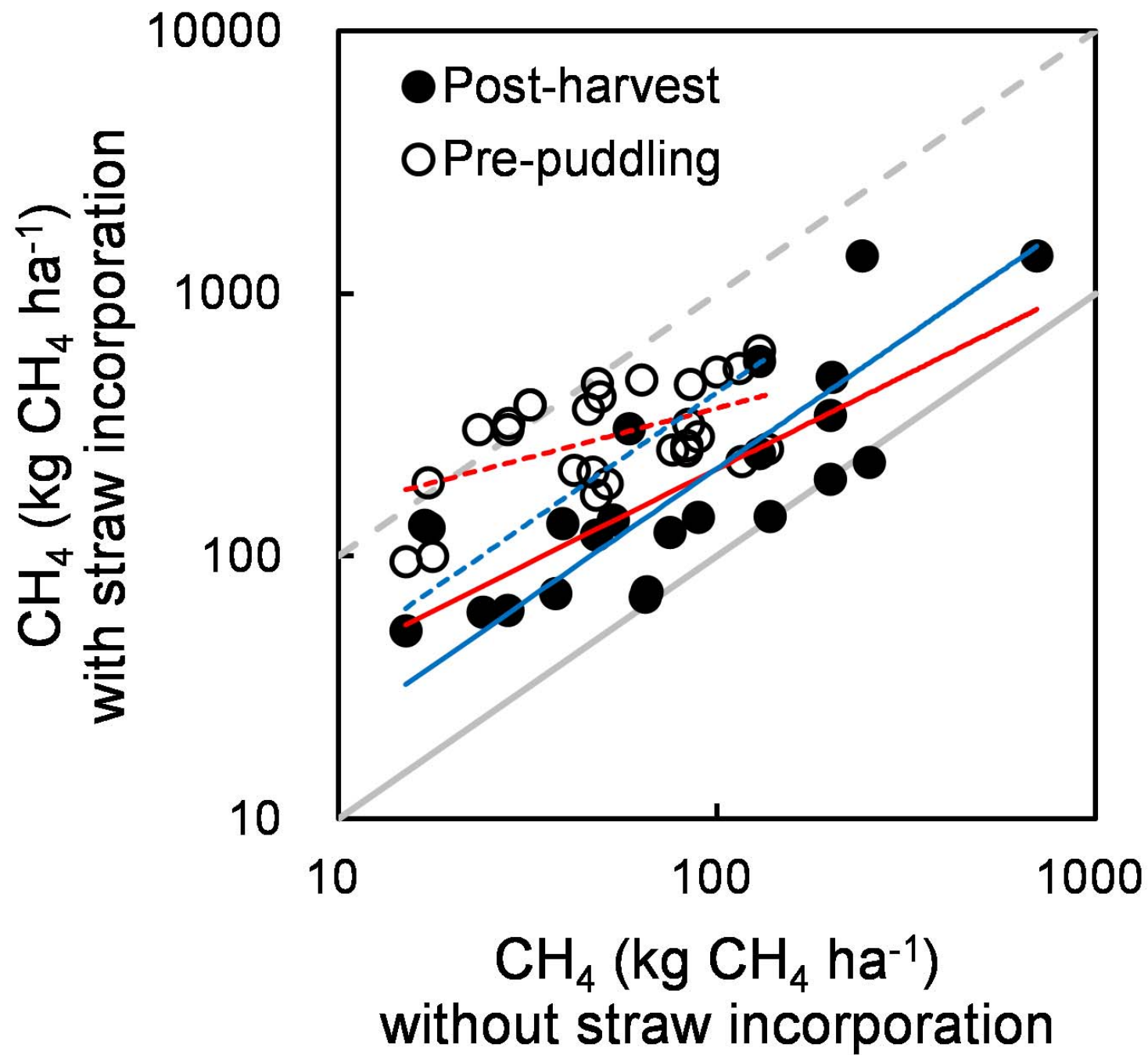
826 Fig. 2. Effect of water drainage on CH₄ emission. The gray solid, broken, and dotted
827 lines show 1:1, 2:1, and 5:1 lines. The black line shows a regression line ($y = 0.61x$)
828 when multiple and single drainage data were compiled.

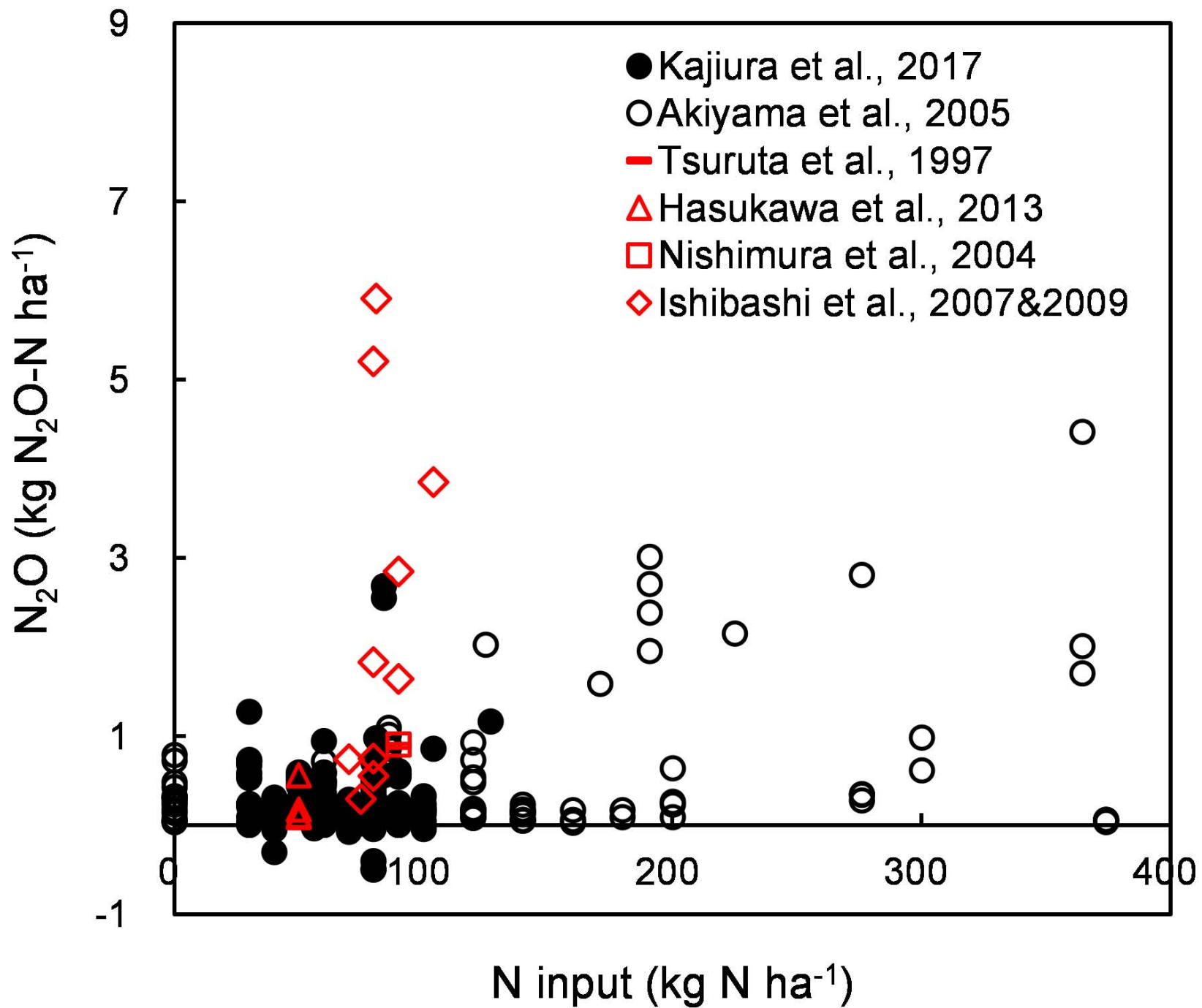
829

830 Fig. 3. N₂O emission as a function of chemical N input in growing season (black
831 symbols) and in fallow season (red symbols). The circled red symbols show high
832 emission obtained in using controlled-release fertilizers, leaving straw on soil surface
833 after harvest, and not conducting tillage (direct seedling).

834

835 Fig. 4. Conceptual model showing the volume changes in methanogenic zone induced
836 by the oxidation of soil organic matter (V_s), rice straw (V_{rs}), and both (V_{s+rs}).
837 Methanogenic zone newly appears by straw incorporation in the soil zones that have not
838 been reduced by the oxidation of soil organic matters to methanogenic level. Thus
839 V_{s+rs} vs V_s relationship can be written as follows: $V_{s+rs} = (1 - V_{rs})V_s + V_{rs}$ (solid
840 line). Without straw incorporation (i.e., $V_{rs} = 0$), $V_{s+rs} = V_s$ (dashed line). The black
841 and gray arrows show the differences in methanogenic-zone volumes ($V_{s+rs} - V_s$) when
842 the initial methanogenic zone was small and large, respectively, before straw
843 incorporation.





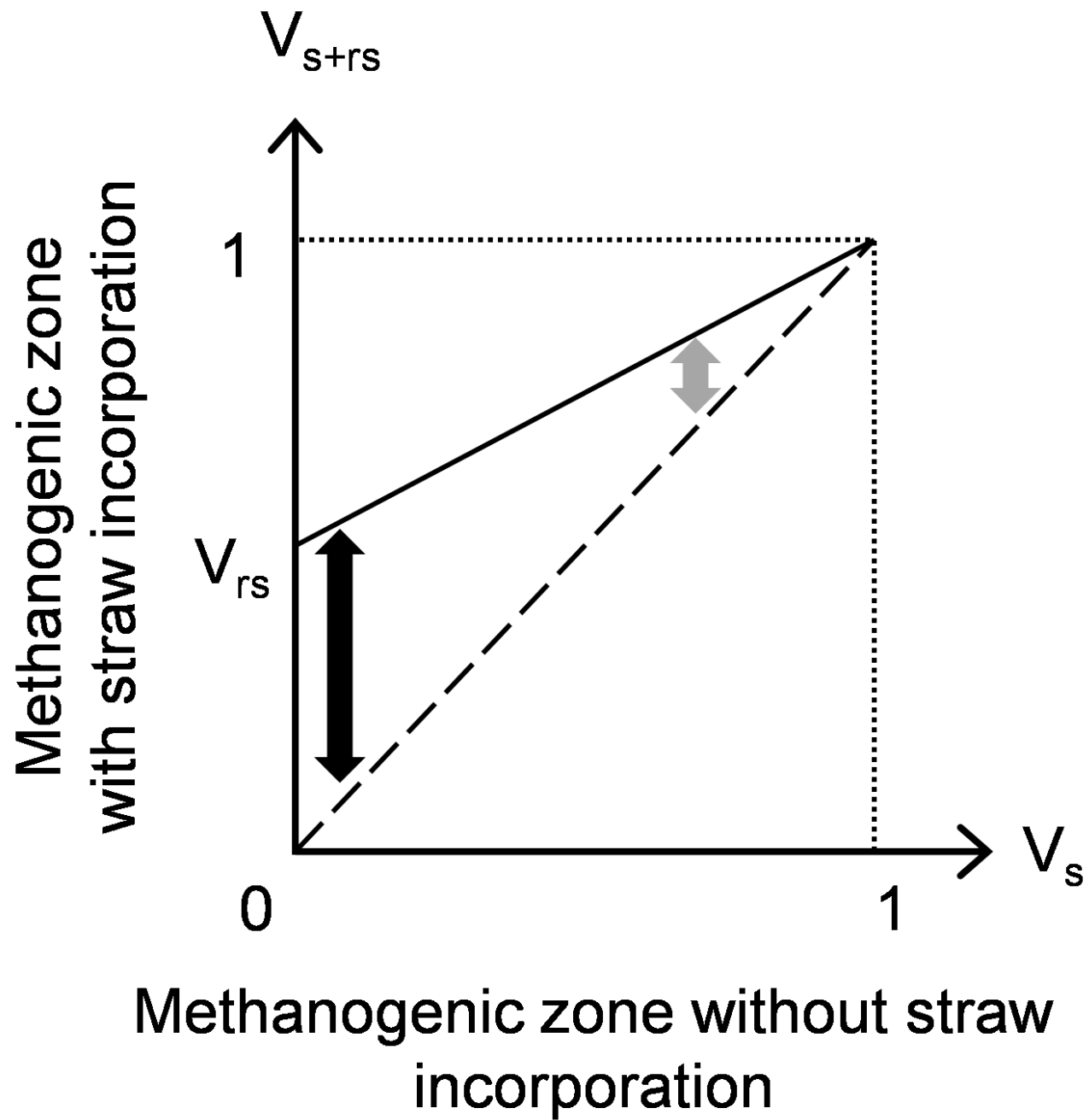


Table A1. Data set on CH4.

Reference	CH ₄ ^a kg CH ₄ ha ⁻¹	Climate zone ^b	Soil Type ^c	Soil pH (H ₂ O)	Soil organic carbon g kg ⁻¹	Water management ^d	Straw incorporation			Manure /compost application	
							Yes/No o	Mass ^e t ha ⁻¹	Timing ^f	Yes/No	
Fumoto <i>et al.</i> , 2008	128	E		5.7	18	CF	Y	6.6	PH	N	
	33	E		5.7	18	CF	N	0.0		N	
	17	E			5.7	18	CF	N	0.0	N	
Goto <i>et al.</i> , 2004	106	N		O	6.1	13	CF	Y	4.7	PH	N
	89	N		O	6.1	13	CF	Y	4.7	PH	N
	52	N		O	6.1	13	CF	Y	4.7	PH	N
	44	N		O	6.1	13	CF	Y	4.7	PH	N
	95	N		O	6.1	13	CF	Y	4.7	PP	N
	80	N		O	6.1	13	CF	Y	4.7	PP	N
	15	N		O	6.1	13	CF	N	0.0		N
	120	N		O	6.1	13	CF	Y	4.7	PH	N
	99	N		O	6.1	13	CF	Y	4.7	PH	N
	171	N		O	6.1	13	CF	Y	4.7	PP	N
	125	N		O	6.1	13	CF	Y	4.7	PP	N
	48	N		O	6.1	13	CF	N	0.0		N
	84	N		O	6.1	13	CF	N	0.0		N
	42	N		O	6.1	13	S	N	0.0		N
	55	N		O	6.1	13	M	N	0.0		N
	54	N		O	6.1	13	CF	N	0.0		N
	31	N		O	6.1	13	S	N	0.0		N
	52	N		O	6.1	13	M	N	0.0		N
	35	N		O	6.1	13	M	N	0.0		N
	89	N		O	6.1	13	CF	N	0.0		N
	47	N		O	6.1	13	S	N	0.0		N
	76	N		O	6.1	13	M	N	0.0		N
	51	N		O	6.1	13	M	N	0.0		N
	286	N		O	6.1	13	CF	Y	3.5	PP	N
	209	N		O	6.1	13	S	Y	3.5	PP	N
	253	N		O	6.1	13	M	Y	3.5	PP	N
	189	N		O	6.1	13	M	Y	3.5	PP	N
	100	N		O	6.1	13	CF	N	0.0		N
	83	N		O	6.1	13	M	N	0.0		N
	503	N		O	6.1	13	CF	Y	3.5	PP	N
251	N		O	6.1	13	S	Y	3.5	PP	N	
444	N		O	6.1	13	M	Y	3.5	PP	N	
263	N		O	6.1	13	M	Y	3.5	PP	N	
51	N		O	6.1	13	M	Y	4.7	PH	N	
43	N		O	6.1	13	M	Y	4.7	PH	N	
Fumoto <i>et al.</i> , 2010	377	N		O	6.6	18	CF	Y	4.0	PP	Y
	251	N		O	6.6	18	S	Y	4.0	PP	Y
	108	N		O	6.6	18	S	Y	4.0	PP	Y
	68	N		O	6.6	18	S	Y	4.0	PP	Y
	341	N		O	6.6	18	CF	Y	4.0	PP	N
	248	N		O	6.6	18	S	Y	4.0	PP	N
	183	N		O	6.6	18	S	Y	4.0	PP	N
116	N		O	6.6	18	S	Y	4.0	PP	N	
Itoh <i>et al.</i> , 2011	358	N		O		29	M	Y		PH	N
	340	N		O		29	M	Y		PH	N
	380	N		O		29	M	Y		PH	N
	241	N		O		29	M	Y		PH	N
	557	N		O		22	M	Y		PH	N
	354	N		O		22	M	Y		PH	N
	469	N		O		22	M	Y		PH	N
	287	N		O		22	M	Y		PH	N
	148	N		O		22	M	Y		PH	N
	180	N		O		22	M	Y		PH	N
	283	N		O		14	S	Y		PH	N
	186	N		O		14	S	Y		PH	N
	118	N		O		14	S	Y		PH	N
	684	N		O		14	S	Y		PH	N
	452	N		O		14	S	Y		PH	N
	329	N		O		14	S	Y		PH	N
	124	E		O		21	M	Y		PH	N
250	E		O		21	M	Y		PH	N	
69	E		O		21	M	Y		PH	N	
307	E		O		21	M	Y		PH	N	

	92	E	O		21	M	Y		PH	N
	82	E	O		21	M	Y		PH	N
	69	E	O		21	M	Y		PH	N
	318	E	O		21	M	Y		PH	N
	36	E	O		15	M	Y		PH	N
	21	E	O		15	M	Y		PH	N
	10	E	O		15	M	Y		PH	N
	84	E	O		15	M	Y		PH	N
	68	E	O		15	M	Y		PH	N
	76	E	O		15	M	Y		PH	N
	525	E	O		12	S	Y		PP	N
	542	E	O		12	S	Y		PP	N
	462	E	O		12	S	Y		PP	N
	662	E	O		12	CF	Y		PP	N
	786	E	O		12	S	Y		PP	N
	717	E	O		12	S	Y		PP	N
	647	E	O		12	S	Y		PP	N
	1044	E	O		12	CF	Y		PP	N
	535	W	O		14	M	Y		PH	N
	233	W	O		14	M	Y		PH	N
	268	W	O		14	M	Y		PH	N
	167	W	O		14	M	Y		PH	N
	67	W	O		14	M	Y		PH	N
	67	W	O		14	M	Y		PH	N
	80	W	O		14	M	Y		PH	N
	54	W	O		14	M	Y		PH	N
	31	W	A		75	S	Y		PH	N
	14	W	A		75	S	Y		PH	N
	29	W	A		75	S	Y		PH	N
	65	W	A		75	CF	Y		PH	N
	61	W	A		75	S	N	0.0		N
	49	W	A		75	S	N	0.0		N
	50	W	A		75	S	N	0.0		N
	52	W	A		75	CF	N	0.0		N
	303	W	O		14	S	N	0.0		N
	335	W	O		14	S	N	0.0		N
	412	W	O		14	S	N	0.0		N
	270	W	O		14	CF	N	0.0		N
	181	W	O		14	S	N	0.0		N
	228	W	O		14	S	N	0.0		N
	155	W	O		14	M	N	0.0		N
Yagi <i>et al.</i> , 1996	148	E	O	6.1	16	CF	Y	5.0	PH	N
	117	E	O	6.1	16	S	Y	5.0	PH	N
	133	E	O	6.1	16	CF	Y	5.0	PH	N
	73	E	O	6.1	16	M	Y	5.0	PH	N
Yagi <i>et al.</i> , 1990	80	E	O			M	N	0.0		N
	82	E	O	5.9	14	M	N	0.0		N
	105	E	O			M	N	0.0		Y
	270	E	O			M	Y	6.0		N
	448	E	H	5.6	34	M	Y	6.0		N
	41	E	A			S	N	0.0		N
	36	E	A	6.5	60	S	N	0.0		N
	59	E	A			S	N	0.0		Y
	98	E	A			S	Y	6.0		N
	126	E	A			S	Y	9.0		N
	6	E	A	5.9	22		N	0.0		N
	11	E	A				Y	6.0		N
Tsuruta <i>et al.</i> , 1997	95	E				S				N
Nishimura <i>et al.</i> , 2004	32	E	O	5.7	19	M	Y	7.0	PH	N
Nishimura <i>et al.</i> , 2008	27	E	O	5.7	19	M	Y	7.4	PH	N
	187	E	O	5.7	20	M	Y	6.8	PH	N
Nishimura <i>et al.</i> , 2011	61	E	O	5.7	19	M	Y	6.8	PH	N
	59	E	O	5.7	19	M	Y	7.0	PH	N
Minamikawa <i>et al.</i> , 2005	12	E	A	6.0	44	CF	N	0.0		N
	41	E	A	6.0	44	CF	N	0.0		N
	76	E	A	6.0	44	CF	N	0.0		N
	59	E	A	6.0	44	CF	N	0.0		N
	79	E	A	6.0	44	CF	N	0.0		N
Shiratori <i>et al.</i> , 2007	233	E	O	5.2	29	M	Y	7.0	PH	N
	136	E	O	5.2	29	M	Y	7.0	PH	N
	76	E	O	5.2	29	M	Y	7.0	PH	N
	32	E	O	5.2	22	M	Y	7.0	PH	N
	40	E	O	5.2	22	M	Y	7.0	PH	N

Ishibashi <i>et al.</i> , 2001	33	E	O	5.2	22	M	Y	7.0	PH	N	
	259	W	O			M	Y	6.0	PH	N	
	2	W	O			M	Y	6.0	SP	Y	
	114	W	O			M	Y	6.0	PH	N	
	22	W	O			M	Y	6.0	SP	Y	
	103	W	O			M	Y	6.0	PH	N	
	79	W	O			M	Y	6.0	SP	Y	
	341	W	O			M	Y	6.0	PP	N	
	164	W	O			M	Y	6.0	SP	N	
	274	W	O			M	Y	6.0	PP	N	
	103	W	O			M	Y	6.0	SP	N	
	221	W	O			M	Y	6.0	PP	N	
	111	W	O			M	Y	6.0	SP	N	
	246	W	O			S	Y	6.0	PP	N	
	138	W	O			S	Y	6.0	SP	N	
	421	W	O			M	Y	8.0	PH	N	
	385	W	O			CF	Y	8.0	SP	N	
	Ishibashi <i>et al.</i> , 2005 and 2009	389	W	O	5.6	30	CF	Y	8.0	SP	N
		393	W	O	5.6	30	CF	Y	8.0	SP	N
		480	W	O	5.6	30	CF	Y	8.0	SP	N
413		W	O	5.6	30	S	Y	8.0	SP	N	
370		W	O	5.6	30	CF	Y	8.0	SP	N	
392		W	O			CF	Y	8.0	PP	N	
354		W	O			CF	Y	8.0	PH	N	
401		W	O			CF	Y	8.0	PH	N	
421		W	O	5.8	28	M	Y	8.0	PH	N	
364		W	O	5.8	28	CF	Y	8.0	PH	N	
449		W	O	5.8	28	CF	Y	8.0	PH	N	
461		W	O	5.8	30	CF	Y	7.7	SP	N	
384		W	O	5.8	30	S	Y	7.7	SP	N	
370		W	O			CF	Y	7.7	PH	N	
Kumagai and Konno, 1998	237	W	O			CF	Y	7.7	PH	N	
	170	N	O	6.3	18	CF	N	0.0		N	
	376	N	O	6.3	14	CF	Y	6.0		N	
	927	N	O			CF	Y	6.0		N	
	432	N	O	6.3	14	CF	Y	6.0		N	
	430	N	O			CF	Y	6.0		N	
	339	N	O	6.3	14	CF	Y	6.0		N	
Hanaki <i>et al.</i> , 2002	615	N	A		31	M	Y	5.0	PH	N	
	136	N	A		35	M	Y	5.0	SP	N	
	578	N	O		17	M	Y	5.0	PH	N	
	211	N	O		13	M	Y	5.0	SP	N	
	115	N	O		29	M	Y	5.0	PH	N	
	183	N	O		29	M	Y	5.0	SP	N	
	354	N	A		31	M	Y	5.0	PH	N	
	178	N	A		35	M	Y	5.0	SP	N	
	335	N	O		17	M	Y	5.0	PH	N	
	163	N	O		13	M	Y	5.0	SP	N	
	104	N	O		29	M	Y	5.0	PH	N	
	48	N	O		29	M	Y	5.0	SP	N	
	Minamikawa <i>et al.</i> , 2010	50	E	O	5.7	19	M	Y	9.9	PH	N
		219	N	A	5.6	83	S	N	0.0		N
Inubushi <i>et al.</i> , 2003	160	N	A	5.6	83	S	N	0.0		N	
	116	N	A	5.6	83	S	N	0.0		N	
	77	N	A	5.6	83	S	N	0.0		N	
Zheng <i>et al.</i> , 2006	93	N	A			S	N	0.0		N	
	87	N	A			S	N	0.0		N	
Naser <i>et al.</i> , 2007	54	N	O	5.8	2	CF	N	0.0		N	
	131	N	O	6.0	3	CF	Y	0.9	PP	N	
	121	N	O	5.8	2	CF	Y	1.2	PP	N	
	519	N	O	5.4	2	CF	Y	2.2	PP	N	
Harada <i>et al.</i> , 2007	544	N	O	5.4	3	CF	Y	2.6	PP	N	
	174	N	O	7.5	19	M	Y	6.4	PP	N	
	109	N	O	7.5	19	M	Y	6.7	SP	N	
	174	N	O	7.5	19	M	Y	6.7	PP	N	
	184	N	O	7.5	19	M	Y	7.3	PP	N	
	95	N	O	7.5	19	M	Y	7.2	SP	N	
	190	N	O	7.5	19	M	Y	7.5	PP	N	
Kumagai <i>et al.</i> , 1993	153	E	O	5.9	14	M	N	0.0		N	
	205	E	O	6.1	16	M	Y	6.0		N	
Minamikawa and Sakai, 2006	73	E	A	6.0	44	M	Y	6.2	PH	N	
	65	E	A	6.0	44	M	N	0.0		N	
	142	E	A	6.0	44	CF	Y	6.2	PH	N	

	138	E	A	6.0	44	CF	N	0.0		N
	70	E	A	6.0	44	M	Y	6.5	PH	N
	65	E	A	6.0	44	M	N	0.0		N
	227	E	A	6.0	44	CF	Y	6.5	PH	N
	252	E	A	6.0	44	CF	N	0.0		N
Hasukawa <i>et al.</i> , 2013	237	W	O	5.5	20	M	Y	9.1	PH	N
	222	W	O	5.5	21	M	Y	9.1	PH	N
	225	W	O	5.5	21	M	Y	9.1	PH	N
	338	W	O	5.6	24	M	Y	9.1	PH	N
Matsumoto <i>et al.</i> , 2002	202	W	O	5.8	41	M	Y	5.0	PH	Y
	241	W	O	5.8	41	M	Y	5.0	PH	Y
	238	W	O	5.8	41	M	Y	5.0	PH	Y
	383	W	O	5.8	41	M	Y	5.0	PH	Y
	107	W	O	5.8	41	M	Y	5.0	PH	Y
	404	W	O	5.8	41	M	Y	5.0	PH	Y
	355	W	O	5.8	41	M	Y	5.0	PH	Y
	253	W	O	5.8	41	M	Y	5.0	PH	Y
Ishibashi <i>et al.</i> , 1997	74	W	O			M	Y		PH	N
	143	W	O			M	Y		PH	N
	265	W	O			M	Y		PH	N
	111	W	O			M	Y		PH	N
	241	W	O			M	Y		PP	N
	361	W	O			CF	Y		PP	N
Kudo <i>et al.</i> , 2014	13	E	A	6.4	23	S	Y	3.5	PH	N
	-1	E	A	6.4	20	M	Y	3.5	PH	N
	14	E	A	6.3	16	M	Y	3.5	PH	N
Riya <i>et al.</i> , 2012	148	E				S	N	0.0		N
	24	E				S	N	0.0		N
	146	E				S	N	0.0		N
	54	E				S	N	0.0		N
	124	E				S	N	0.0		Y
	61	E				S	N	0.0		Y
	95	E				S	N	0.0		Y
	79	E				S	N	0.0		Y
Sugii <i>et al.</i> , 1999	340	E			3	CF	N	0.0		N
	597	E			4	CF	N	0.0		N
	391	E			24	CF	N	0.0		N
	459	E			23	CF	N	0.0		N
	301	E			19	CF	N	0.0		N
	375	E			14	CF	N	0.0		N
	283	E	H		220	CF	N	0.0		N
Uoki and Noda, 2001	969	E	H		195	CF	N	0.0		N
	188	W	O		13	S				N
	98	W	O		23	S				N
	85	W	O		16	S				N
	310	W	O		13	S				N
	158	W	O		23	S				N
	64	W	O		16	S				N
	113	W	O		13	S				N
	59	W	O		23	S				N
	37	W	O		16	S				N
	355	W	O		13	S	Y	6.0	PP	N
	267	W	O		13	M	Y	6.0	PP	N
	226	W	O		13	M	Y	6.0	PP	N
	156	W	O		13	S	Y	6.0	PH	N
	101	W	O		13	M	Y	6.0	PH	N
	33	W	O		13	M	Y	6.0	PH	N
	147	W	O		13	S	Y	6.0	PH	N
	121	W	O		13	M	Y	6.0	PH	N
	38	W	O		13	M	Y	6.0	PH	N
	98	W	O		13	M	Y	6.0	PH	N
	48	W	O		13	M	Y	6.0	PH	N
	157	W	O		13	M	Y	6.0	PH	N
	50	W	O		13	M	Y	6.0	PH	N
Akai <i>et al.</i> , 1996	362	W	O			M	Y			N
	48	W	O			M	Y			N
	385	W	O			M	Y			N
	124	W	O			M	Y			N
	678	W	O			M	Y			N
	28	W	O			M	Y			N
Koshiba and Kato, 1995	95	E	A	6.2	34	M	N	0.0		N
	66	E	A	6.2	34	M	N	0.0		N
	71	E	A	6.2	34	M	N	0.0		N

	517	E	O	6.7	30	M	Y	5.0		N
	42	E	O	6.7	30	M	N	0.0		Y
	129	E	O	6.7	30	M	N	0.0		N
Shinoda <i>et al.</i> , 1999	201	E			4	CF	N	0.0		N
	478	E			6	CF	Y	5.0	PH	N
	699	E			32	CF	N	0.0		N
	1390	E			34	CF	Y	5.0	PH	N
	129	E			17	CF	N	0.0		N
	553	E			20	CF	Y	5.0	PH	N
	241	E	H		314	CF	N	0.0		N
	1393	E	H		303	CF	Y	5.0	PH	N
	605	E	O			S	Y	5.0	PP	N
	245	E	O		23	S	Y	5.0	PH	N
	206	E	O		22	S	N	0.0		Y
	129	E	O		22	S	N	0.0		N
	76	E	O		23	S	Y	5.0	PH	N
	73	E	O			S	Y	5.0	PH	N
	74	E	O			S	Y	5.0	PH	N
	59	E	O		24	S	N	0.0		N
	59	E	O			M	N	0.0		N
	52	E	O			M	N	0.0		N
Kouzai and Hiraki, 1996	214	W	O	5.4	13	M	N	0.0		N
	351	W	O	5.3	15	M	Y	6.0	PH	N
	829	W	O	5.2	15	M	Y	5.0	PP	N
	360	W	O	5.5	13	M	N	0.0		N
	434	W	O	5.6	14	M	Y	6.0	PH	N
	1128	W	O	5.3	15	M	Y	5.0	PP	N
	424	W	O	5.2	12	M	N	0.0		N
	593	W	O	5.1	14	M	Y	6.0	PH	N
	1189	W	O	4.9	14	M	Y	5.0	PP	N
Sasaki <i>et al.</i> , 2002	163	W	O	6.3	19	S				N
	109	W	O	6.3	19	S				N
	122	W	O	6.3	19	S				N
	60	W	O	5.3	18	S	N	0.0		N
	60	W	O	5.3	18	S	N	0.0		N
	80	W			33	CF	N	0.0		N
	89	W			33	CF	N	0.0		N
	146	W			33	CF	N	0.0		N
	126	W			33	CF	Y		SP	N
	152	W			33	CF	Y		SP	N
	160	W			33	CF	Y		SP	N
Suzuki, 1995	53	E	A	6.6	87	M	N	0.0		N
	38	E	A	6.6	91	M	N	0.0		Y
	137	E	A	6.5	87	M	Y	5.0	PH	N
	24	E	A	6.6	87	M	N	0.0		N
	19	E	A	6.6	91	M	N	0.0		Y
	61	E	A	6.5	87	M	Y	5.0	PH	N
	89	E	A	6.6	87	M	N	0.0		N
	65	E	A	6.6	91	M	N	0.0		Y
	140	E	A	6.5	87	M	Y	5.0	PH	N
	164	E	A	6.6	89	M	Y	5.0	PH	N
	174	E	A	6.7	70	M	N	0.0		Y
	206	E	A	6.8	73	M	N	0.0		Y
	119	E	A	6.7	70	M	N	0.0		Y
	78	E	A	6.8	73	M	N	0.0		Y
	73	E	O	5.8	17	M	N	0.0		N
	111	E	O	6.3	23	M	N	0.0		Y
	745	E	O	6.0	20	M	Y	10.7	PP	N
	43	E	O	5.8	17	M	N	0.0		N
	106	E	O	6.3	23	M	N	0.0		Y
	273	E	O	6.0	20	M	Y	10.7	PP	N
	60	E	O	5.8	17	M	N	0.0		N
	116	E	O	6.3	23	M	N	0.0		Y
	369	E	O	6.0	20	M	Y	10.7	PP	N
	429	E	O	6.0	19	M	Y	9.1	PP	N
Kumagai, 2002	252	N	O	5.3	13	M	Y	6.0	PP	N
	83	N	O	5.4	12	M	N	0.0		N
	230	N	O	5.3	13	M	Y	6.0	PP	N
	116	N	O	5.4	12	M	N	0.0		N
	212	N	O	5.3	13	M	Y	6.0	PP	N
	42	N	O	5.4	12	M	N	0.0		N
	261	N	O	5.6	12	M	N	0.0		Y
	69	N	A	5.7	60	M	N	0.0		Y

	21	N	A	5.4	69	M	N	0.0		N
	94	N	A	5.7	60	M	N	0.0		Y
	56	N	A	5.4	69	M	N	0.0		N
	118	N	A	5.7	60	M	N	0.0		Y
	59	N	A	5.4	69	M	N	0.0		N
Miura, 2003	59	N	O	5.5	13	S	N	0.0		N
	305	N	O	6.3	16	S	Y	6.0	PH	N
	37	N	O	6.2	15	S	N	0.0		N
	72	N	O	6.2	15	S	Y	6.0	PH	N
	46	N	O	6.1	33	CF	N	0.0		N
	364	N	O	6.1	33	CF	Y	6.0	PP	N
	17	N	O	5.5	13	S	N	0.0		N
	131	N	O	6.3	16	S	Y	6.0	PH	N
	28	N	O	6.2	15	CF	N	0.0		N
	320	N	O	6.2	15	CF	Y	6.0	PP	N
	203	N	O	6.2	15	CF	N	0.0		N
	18	N	A	5.8	52	CF	N	0.0		N
	100	N	A	5.8	52	CF	Y	6.0	PP	N
	39	N	O	5.5	13	S	N	0.0		N
	133	N	O	6.3	16	S	Y	6.0	PH	N
	75	N	O	6.2	15	S	N	0.0		N
	123	N	O	6.2	15	S	Y	6.0	PH	N
	28	N	O	6.2	15	CF	N	0.0		N
	34	N	O	6.2	15	CF	Y	6.0	PH	N
	62	N	O	6.2	15	CF	Y	6.0	PH	N
	109	N	O	6.2	15	CF	Y	6.0	PP	N
	303	N	O	6.2	15	CF	Y	6.0	PP	N
Miyoshi et al., 1994	29	W	O	5.6	14	M	N	0.0		Y
	62	W	O	5.9	20	M	N	0.0		Y
	103	W	O	6.4	31	M	N	0.0		Y
Hayashi et al., 2010	186	N	O				Y	6.3	PP	N
	241	N	O				Y	6.3	PP	Y
	166	N	O				Y	6.3	PP	Y
	160	N	O				Y	6.3	PP	Y
	251	N	O				Y	5.0	PP	N
	417	N	O				Y	5.0	PP	Y
	252	N	O				Y	5.0	PP	Y
	255	N	O				Y	5.0	PP	Y
	167	N	O				Y	3.7	PP	N
	204	N	O				Y	3.7	PP	Y
	208	N	O				Y	3.7	PP	Y
	254	N	O				Y	3.7	PP	Y
	268	N	O	6.2	29	M	N	0.0		Y
	194	N	O	6.2	29	M	N	0.0		Y
	522	N	O	5.8	28	M	Y	6.0	PP	N
	244	N	O	6.5	28	M	N	0.0		Y
	488	N	O	6.5	28	M	Y	6.0	PP	N
	356	N	O	6.2	29	M	N	0.0		Y
	42	N	O	6.2	29	M	N	0.0		Y
	295	N	O	5.8	28	M	Y	6.0	PP	N
	394	N	O	6.5	24	M	N	0.0		Y
	547	N	O	6.5	24	M	Y	6.0	PP	N
	225	N	O	5.6	22	M	N	0.0		Y
	214	N	O	5.6	22	M	N	0.0		Y
	274	N	O	5.6	22	M	N	0.0		Y
Miura and Kanno, 1993	93	N	O	5.8	12	S	Y		PH	N
	83	N	O	5.8	16	S	Y		PH	N
	136	N	O	6.0	15	S	Y		PH	N
		N	O	5.8	12		Y		PH	N
		N	O	6.0	15		Y		PH	N
Ueki et al., 1995	199	N	O			S	N	0.0		N
	209	N	O			S	N	0.0		N
	83	N	O			S	N	0.0		Y
	237	N	O			S	N	0.0		Y
	113	N	O			S	N	0.0		Y
	114	N	O			S	N	0.0		Y
	196	N	O			S	Y	5.8	PH	N
	202	N	O			S	Y	5.8	PH	N
	343	N	O			S	Y	8.8	PH	N
	292	N	O			S	Y	8.8	PH	N
Omori et al., 2013	331	W				S	N	0.0		N
	252	W				S	N	0.0		N
	293	W				S	N	0.0		Y

	318	W				S	N	0.0		Y
Maruyama <i>et al.</i> , 2012	65	W				S				N
	24	W				S				N
Ichikawa <i>et al.</i> , 2011	255	N				S	Y	6.0	PP	N
	196	N				S	N	0.0		Y
	136	N				S	N	0.0		N
Nitta <i>et al.</i> , 2009	222	W				CF	N	0.0		Y
	190	W				CF	N	0.0		Y
	189	W				CF	N	0.0		Y
	157	W				CF	N	0.0		N
	149	W				CF	N	0.0		N
Kamio and Kobayashi, 2004	279	N				M	Y	6.0		N
	440	N				M	Y	6.0		N
	503	N				M	Y	6.0		N
	496	N				M	N	0.0		N
	329	N				M	N	0.0		N
Kamio and Kotabe, 2002	278	N		4.5	22	M	Y	6.0		N
	1085	N		5.0	20	CF	N	0.0		N
Kamio and Kaino, 2003	291	N			17	M	Y	6.0		N
	143	N			17	M	Y	6.0		N
	168	N			41	M	Y	6.0		N
	89	N			41	M	Y	6.0		N
	100	E	A	6.5	71	CF	Y	6.0	PH	N
	99	E	A	6.5	71	CF	Y	6.0	PH	N
	56	E	A	6.5	71	CF	Y	6.0	PH	N
Kudo <i>et al.</i> , 2012	10	E	A	6.3	11	S	N	0.0		N
	2	E	A	6.2	11	M	N	0.0		N
	-1	E	A	6.2	9	M	N	0.0		N
Iida <i>et al.</i> , 2007	555	N								N
	788	N		6.4						N
Minami and Yagi, 1988	9	E	O	6.2	25	CF				N
	1	E	A	5.9	22	CF				N
Horikawa <i>et al.</i> , 2007	273	N	O		31	CF				N
	255	N	O		31	CF				N
	303	N	O		31	CF				N
	338	N	O		31	CF				N
Hosono and Nouchi, 1996	114	E	O	6.0	19	S	N	0.0		N
	32	E	O	6.0	19	CF	N	0.0		N
	85	E	O	6.0	19	CF	N	0.0		N
	84	E	O	6.0	19	CF	N	0.0		N
	519	E	O	6.0	19	S	Y	7.0	PP	N
	378	E	O	6.0	19	CF	Y	7.0	PP	N
	451	E	O	6.0	19	CF	Y	7.0	PP	N
	318	E	O	6.0	19	CF	Y	7.0	PP	N
Nouchi <i>et al.</i> , 1994	32	E	O	6.0	19	CF	N	0.0		N
	378	E	O	6.0	19	CF	Y	7.0	PP	N
	494	E	O	6.0	19	CF	Y	7.0	PP	N
Watanabe <i>et al.</i> , 1995	174	E		5.7	12	S				N
	159	E		5.7	12	S				N
	185	E		5.7	12	S				N
	207	E		5.7	12	S				N
	154	E		5.7	12	S				N
	207	E		5.7	12	S				N
	196	E		5.7	12	S				N
	134	E		5.7	12	S				N
Shiono <i>et al.</i> , 2014	72	N	O	5.9	24	M	N	0.0		N
	190	N	O	5.9	24	M	Y	5.6	PP	N
	404	N	O	5.6	25	M	Y	6.0	PP	N
	468	N	O	5.6	25	M	Y	6.0	PP	N
	455	N	O	5.9	25	M	Y	6.0	PP	N
	302	N	O	5.9	25	M	Y	6.0	PP	N
	95	N	O	5.9	22	M	N	0.0		N
	17	N	O	5.9	22	M	N	0.0		N
	49	N	O	5.6	23	M	N	0.0		N
	63	N	O	5.6	23	M	N	0.0		N
	48	N	O	5.7	22	M	N	0.0		N
	23	N	O	5.7	22	M	N	0.0		N
Win <i>et al.</i> , 2014	19	E	O		30	CF	N	0.0		N
	2	E	O		29	CF	N	0.0		N
	9	E	O		29	CF	N	0.0		Y
	2	E	O		28	CF	N	0.0		Y
	116	E	O		31	CF	N	0.0		N
	100	E	O		31	CF	N	0.0		N

	213	E	O		33	CF	N	0.0		Y
	184	E	O		33	CF	N	0.0		Y
	362	E	O		33	CF	N	0.0		N
	343	E	O		34	CF	N	0.0		N
	395	E	O		34	CF	N	0.0		Y
	456	E	O		34	CF	N	0.0		Y
	411	E	O		33	CF	N	0.0		N
	330	E	O		35	CF	N	0.0		N
	652	E	O		36	CF	N	0.0		Y
Eusufzai et al, 2010	1090	E	O		35	CF	N	0.0		Y
	4	N	A				N	0.0		N
	17	N	A				N	0.0		N
	61	N	A				N	0.0		N
Naser et al., 2005	1136	N	H		66	S	Y	8.3		N
Toma et al., 2013	51	W		6.4	15	M	Y	2.1	PH	N
	888	W		6.3	18	M	N	0.0		Y
Tsuruta, 1997	100		A			M	Y	6.0	PH	N
	59		A			M	Y	6.0	PH	N
	96		A			M	Y	6.0	PH	N
	178		O			M	Y	6.0	PH	N
	295		O			M	Y	6.0	PH	N
	159		O			M	Y	6.0	PH	N
	199		O			M	Y	6.0	PH	N
	193		O			M	Y	6.0	PH	N
	180		O			M	Y	6.0	PH	N
	163		O			M	Y	6.0	PH	N
	204		O			M	Y	6.0	PH	N
	155		O			M	Y	6.0	PH	N
	163		H			M	Y	6.0	PH	N
	226		H			M	Y	6.0	PH	N
	415		H			M	Y	6.0	PH	N
	37		A			M	N	0.0		N
	64		A			M	N	0.0		N
	97		A			M	N	0.0		N
	38		H			M	N	0.0		N
	78		H			M	N	0.0		N
	188		H			M	N	0.0		N
	111		O			M	N	0.0		N
	105		O			M	N	0.0		N
	132		O			M	N	0.0		N
	117		O			M	N	0.0		N
	109		O			M	N	0.0		N
	143		O			M	N	0.0		N
	83		O			M	N	0.0		N
	110		O			M	N	0.0		N
	130		O			M	N	0.0		N
	59		A			M	Y	6.0	PH	N
	100		A			M	Y	6.0	PH	N
	96		A			M	Y	6.0	PH	N
	108		H			M	Y	6.0	PH	N
	152		H			M	Y	6.0	PH	N
	274		H			M	Y	6.0	PH	N
	177		O			M	Y	6.0	PH	N
	159		O			M	Y	6.0	PH	N
	294		O			M	Y	6.0	PH	N
	197		O			M	Y	6.0	PH	N
	198		O			M	Y	6.0	PH	N
	188		O			M	Y	6.0	PH	N
	174		O			M	Y	6.0	PH	N
	153		O			M	Y	6.0	PH	N
	208		O			M	Y	6.0	PH	N
Oishi et al., 1994	91	W	O		17	M	Y	7.5	PH	N
Kitada et al., 1993	462	E	O	6.2	25	M	Y	6.0	PP	N
	356	E	O	6.2	25	M	Y	6.0	PP	N
	3	E	O	6.2	25	M	N	0.0		N
	71	E	A	5.9	22	M	Y	6.0	PP	N
Hatanaka et al., 1999	146	N	H			CF	Y		SP	N
	44	N	H			M	Y		SP	N
	249	N	H			M	Y		PH	N
	402	N	H			CF	Y		PH	N

a. When numerical data was not available, figures were digitized by GSYS2.4

b. N: north; E: east; W: west

c. A: Andosols; H: Histosols; O: other soil types

d. CF: continuous flooding; S: single drainage (midsummer drainage); M: multiple drainages (intermittent drainage with/without midsummer drainage)

e. Fresh weight. Dry weight was converted to fresh weight by multiplying 1.15.

f. PP: pre-puddling; PH: post-harvest; SP: spreading (no-tillage)

Table A2. Data set on N₂O.

Reference	N ₂ O ^a		Water management	Chemical N input	Manure/compost application	
	Growing season	Fallow season				
	Annual					Yes/No
	kg N ₂ O ha ⁻¹			kg N ha ⁻¹		
Fumoto <i>et al.</i> , 2010	0.29		CF	60	Y	
	0.30		S	60	Y	
	0.26		S	60	Y	
	0.28		S	60	Y	
	0.36		CF	60	N	
	0.58		S	60	N	
	0.44		S	60	N	
	0.48		S	60	N	
	Itoh <i>et al.</i> , 2011	0.00		M	80	N
		0.00		M	80	N
0.00			M	80	N	
0.00			M	80	N	
0.15			M	60	N	
0.14			M	60	N	
0.00			M	60	N	
0.94			M	60	N	
0.00			M	60	N	
0.00			M	60	N	
0.05			S	60	N	
0.00			S	60	N	
0.50			S	60	N	
0.12			S	60	N	
0.26			S	60	N	
0.21			S	60	N	
0.00			M	30	N	
0.21			M	30	N	
0.00			M	30	N	
0.00			M	30	N	
0.24			M	30	N	
0.53			M	30	N	
0.71			M	30	N	
0.71			M	30	N	
0.00			M	80	N	
0.32			M	80	N	
0.26			M	80	N	
0.14			M	80	N	
0.18			M	80	N	
0.36			M	80	N	
0.12			S	56	N	
0.10			S	56	N	
0.03			S	56	N	
0.15			CF	56	N	
0.02		S	56	N		
0.12		S	56	N		
-0.03		S	56	N		
0.35		CF	56	N		
0.37		M	59	N		
0.15		M	59	N		
0.35		M	59	N		
0.15		M	59	N		
0.19		M	59	N		
0.00		M	59	N		
0.00		M	59	N		
0.06		M	59	N		
0.22		S	90	N		
0.58		S	90	N		
0.58		S	90	N		
0.54		CF	90	N		
0.00		S	90	N		
0.00		S	90	N		
0.00		S	90	N		

	0.00			CF	90	N
	0.00			S	70	N
	0.06			S	70	N
	0.28			S	70	N
	0.06			CF	70	N
	0.20			S	70	N
	0.08			S	70	N
	-0.08			M	70	N
Yagi <i>et al.</i> , 1996	0.16			CF	90	N
	0.15			M	90	N
Tsuruta <i>et al.</i> , 1997		0.89		S	90	N
Nishimura <i>et al.</i> , 2004	0.04	0.91	0.95	M	90	N
Nishimura <i>et al.</i> , 2011			0.86	M	90	N
			0.64	M	90	N
Ishibashi <i>et al.</i> , 2007 and 2009	0.18	0.55	0.74	CF	80	N
	0.15	2.85	3.00	CF	90	N
	0.20	5.91	6.11	CF	81	N
	0.86	3.85	4.71	S	104	N
	1.16			CF	127	N
	0.70	0.75	1.45	CF	80	N
	0.61	1.64	2.25	CF	90	N
	0.97			CF	81	N
	0.18	0.74	0.92	M	70	N
	0.02	0.29	0.31	CF	75	N
	0.11		0.11	CF	83	N
	-0.05	5.21	5.16	CF	80	N
	0.00	1.83	1.83	CF	80	N
Minamikawa <i>et al.</i> , 2010			0.62	M	90	N
Harada <i>et al.</i> , 2007	0.16			M	50	N
	0.26			M	50	N
	0.28			M	50	N
Hasukawa <i>et al.</i> , 2013	0.44	0.10	0.38	M	50	N
	0.59	0.56	0.94	M	50	N
	0.55	0.18	0.53	M	50	N
	0.50	0.09	0.41	M	50	N
Kudo <i>et al.</i> , 2014	0.68			S	30	N
	1.27			M	30	N
	0.58			M	30	N
Riya <i>et al.</i> , 2012	0.56			S	84	N
	2.55			S	84	N
	0.19			S	84	N
	2.68			S	84	N
Hayashi <i>et al.</i> , 2010	0.12				50	N
	0.15				40	Y
	0.08				40	Y
	0.12				40	Y
	0.24				40	N
	0.22				40	Y
	0.31				40	Y
	0.19				40	Y
	0.28				40	N
	0.16				40	Y
	0.07				40	Y
	-0.05				40	Y
Miura and Kanno, 1994	0.13			S	100	N
	0.23			S	100	N
	0.32			S	100	N
	0.07			S	100	N
	0.07			S	100	N
	-0.10			S	100	N
	0.02			S	100	N
	0.02			S	100	N
Kudo <i>et al.</i> , 2012	0.03			S	30	N
	0.09			M	30	N
	0.73			M	30	N
Iida <i>et al.</i> , 2007	0.10					N
	0.33					N
Shiono <i>et al.</i> , 2014	0.30			M	40	N
	0.10			M	80	N
	-0.50			M	80	N
	0.20			M	80	N
	-0.50			M	80	N
	0.40			M	80	N

	-0.30		M	40	N
	0.50		M	80	N
	-0.40		M	80	N
	0.30		M	80	N
	0.10		M	80	N
	0.20		M	80	N
Naser <i>et al.</i> , 2005		0.79	S	28	N
Toma <i>et al.</i> , 2013	0.24		M	80	N
	0.35		M	0	Y

a. When numerical data was not available, figures were digitized by GSYS2.4

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-Supplementary information-

Methane and nitrous oxide emissions from paddy fields in Japan: An assessment of
controlling factor using an intensive regional data set

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Effects of soil properties on CH₄

We examined the contribution of soil properties—soil organic matter content (Total carbon: TC) and pH—on the CH₄ by fitting Eq. S1 to collected data using R ver. 3.1.2. The data was a part of those used in the section 3.1.1 (n = 221).

$$\ln(\text{FluxCH}_4) = E1_i + E2_j + \ln(E3) + E4_n + M1_k + M2_l + M3_m \quad (\text{Eq. S1})$$

Table S1. Parameters and variables for Eq. S1.

Parameters	Symbol in the Eq.	Specific categorical variables used
TC content	$E3$	(continuous variable)
pH	$E4_n$	5.0–5.5, 5.5–6.0, 6.0–6.5, 6.5–7.0, 7.5–8.0

Table S2. Effects of soil pH and TC content on CH₄.

Parameters	Estimate	SD	t	P	Lower 95% CL*	Upper 95% CL*	Response ratio		
							mean	Lower 95% CL*	Upper 95% CL*
Intercept	2.67	0.36	7.33	0.000	1.95	3.38			
Environment									
Climate zone									
North	0.14	0.13	1.08	0.280	-0.12	0.41	1.15	0.89	1.50
East	0.00						1.00		
West	0.48	0.23	2.05	0.041	0.02	0.93	1.61	1.02	2.54
Soil type									
Others	0.00	0.00					1.00		
Andosols	-0.65	0.19	-3.43	0.001	-1.02	-0.28	0.52	0.36	0.76
Histosols	na	na	na	na	na	na	na	na	na
ln TC	0.51	0.12	4.17	0.000	0.27	0.75			
Soil pH									
5.0-5.5	0.06	0.19	0.29	0.769	-0.32	0.43	1.06	0.73	1.54
5.5-6.0	0.07	0.14	0.51	0.612	-0.21	0.35	1.07	0.81	1.42
6.0-6.5	0.00						1.00		
6.5-7.0	-0.08	0.18	-0.46	0.646	-0.44	0.27	0.92	0.65	1.31
7.5-8.0	-0.44	0.34	-1.31	0.190	-1.11	0.22	0.64	0.33	1.25
Management									
Straw incorporation									
None	0.00						1.00		

Post-harvest	0.42	0.13	3.23	0.001	0.16	0.68	1.52	1.18	1.97
Spreading (No tillage)	1.10	0.32	3.47	0.001	0.48	1.73	3.02	1.61	5.64
Pre-puddling	1.48	0.14	10.60	0.000	1.21	1.76	4.40	3.34	5.80
Manure/Compost application									
No	0.00						1.00		
Yes	0.38	0.16	2.30	0.022	0.05	0.70	1.46	1.06	2.01
Water management									
Continuous flooding	0.00						1.00		
Single/Multiple drainage	-0.20	0.12	-1.60	0.112	-0.44	0.05	0.82	0.64	1.05

*CL: confidence limit

Figures

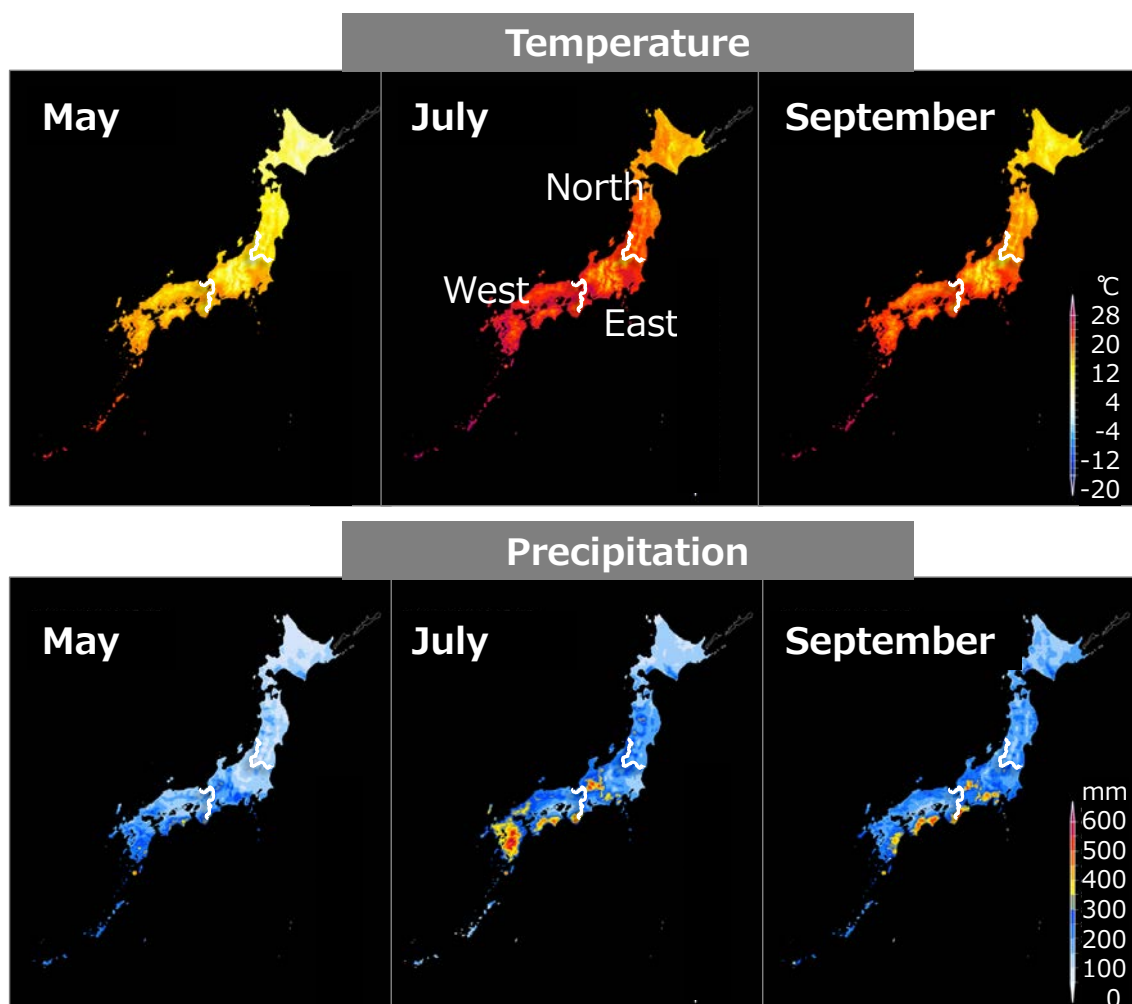


Fig. S1. Monthly averaged precipitation and temperature in growing season (Japan Meteorology Agency, <http://www.data.jma.go.jp/obd/stats/etrn/view/atlas.html>). Darker

and warmer color show higher temperature and larger precipitation, respectively. Climate zone was separated by white lines.

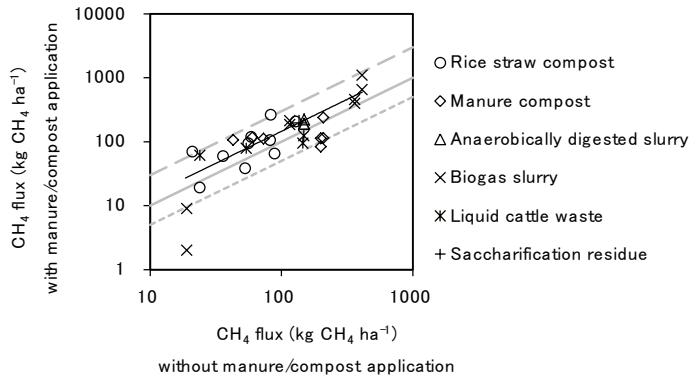


Fig. S2. Effect of manure/compost application on CH₄ emission. The gray dotted, solid, and broken lines show 1:0.5, 1:1, and 1:3 lines. The black solid line is a regression line ($y = 1.46x$) when all data were compiled.

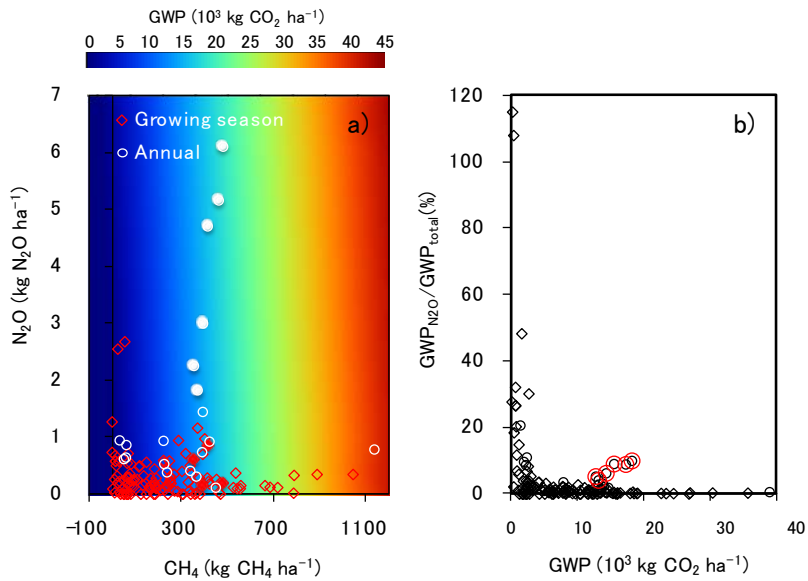


Fig. S3. N₂O contribution to total GWP. Closed symbols in panel a and red circled symbol in panel b were extremely high N₂O data obtained by Ishibashi et al., 2007 and 2009.